Organohalogen Residues of Fishes from Different Trophic Levels

A thesis submitted to the University of Dhaka, Bangladesh in the fulfillment of requirements for the degree of Doctor of Philosophy

Submitted by

Tonima Mustafa

Registration Number: 26/2012-2013 Session: 2012-2013

Department of Zoology University of Dhaka Bangladesh

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Dedicated

To

My beloved parents Late A. K. M. Golam Mustafa and Tahura Begum

and

My husband Engr. Md. Anamul Haque Bhuiyan

CERTIFICATION

The dissertation entitled **"Organohalogen Residues of Fishes from Different Trophic Levels"** submitted to the Department of Zoology, Faculty of Biological Sciences, University of Dhaka, Bangladesh in partial fulfillment of the requirements for the degree of Doctor of Philosophy. We certified that the candidate, **Tonima Mustafa** (Registration No. 26/2012-2013) has been completed her research under our supervision and suggestions. We have read this dissertation and that, in our opinion, it is fully adequate in scope and quality as a dissertation for the degree of Doctor of Philosophy. Experimental work described in the thesis has been carried out by the author at the Department of Chemistry, University of Dhaka. The work has not been and will not be presented for any other degree. It is further certified that to the best of our knowledge the thesis contains original research.

Supervisors

Professor Dr. M. Niamul Naser Professor Dr. Gulshan Ara Latifa Department of Zoology Department of Zoology University of Dhaka University of Dhaka Dhaka Dhaka

Professor Dr. Nilufar Nahar Department of Chemistry University of Dhaka Dhaka

.

DECLARATION

I do hereby declare that the research work entitled "**Organohalogen Residues of Fishes from Different Trophic Levels"** submitted to the Department of Zoology, Faculty of Biological Sciences, University of Dhaka, Dhaka, Bangladesh, for the degree of Doctor of Philosophy is the result of my own analysis. The thesis or part of it has not been presented before for any other degree.

Tonima Mustafa Ph.D. Student Registration No.: 26/2012-13 Department of Zoology University of Dhaka Dhaka

Author

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ABSTRACT

Global contexts on human health hazards through pesticide residues become a serious focus and environmental issues today. Fishes are used extensively for environmental monitoring because they uptake contaminants directly from water and food. Generally the ability of the fish to metabolize organohalogen is moderate, thus contaminants load in fish are well reflective of the state of pollution in surrounding environments. The present study was conducted to assess the concentrations and patterns of organohalogen pesticide residues i.e. DDT and its metabolites (4,4́-DDT, 2,4́- DDT, DDD and DDE) in fishes and prawn species of different trophic levels of four seasons from Sonargaon Upazila of the Meghna River. The samples were collected between the periods of 2015-2016. The number of species available in the seasons varied from twenty-two to twenty-four i.e. rainy season (twenty-four), autumn (twenty-two), winter (twenty-three) and summer (twenty-two). Analysis of the samples for DDTs residues were carried out using Gas Chromatograph with Electron Capture Detector (GC-ECD). The samples were extracted by Quick, Easy, Cheap, Effective, Rugged and Safe (QuEChERS) method and the sample extracts were cleaned-up by using $H₂SO₄$ Linearities $(r²)$ were > 0.9950 for calibrations. The recoveries were 88.67% 104.89% (20 ng g⁻¹), 70.10% - 101.32% (10 ng g⁻¹) and 71.64% - 113.83% (5 ng g⁻¹). The limit of detection was found 0.0625 ng $g⁻¹$ in fish samples. The concentrations of total DDTs residue in fish and prawn tissues of rainy-season, autumn, winter and summer varied between 2.64 \pm 0.35 ng g⁻¹ to 191.14 \pm 31.18 ng g⁻¹, 16.42 \pm 1.90 ng g⁻¹ ¹ to 271.50 \pm 6.17 ng g⁻¹, 3.88 \pm 0.60 ng g⁻¹ to 141.57 \pm 10.24 ng g⁻¹, 157.58 \pm 1.15 ng g⁻¹ to 1660.89 \pm 157.9 ng g⁻¹ wet weight (ww) respectively. The year round highest concentrations were observed in Bacha (*Eutropiichthys vacha*). However the lowest levels of total DDTs were observed in Kachki (*Corica soborna*) in rainy-season and summer, Khalisha (*Trychogaster fasciata*) in autumn and Goldachingri (*Macrobrachium rosenbergii*) in winter. Considering the average concentrations of total DDTs residue of four seasons, the twenty fish and prawn species that analysed in all seasons showed the chronology of Kachki (*Corica soborna*) < Chanda (*Parambassis ranga*) < Shing (*Heteropnuestes fossilis)* < Ghainna *(Labeo gonius*) < Rui (*Labeo rohita*) < *Systomus sarana* (Sharpunti) < Bata (*Cirrhinus reba*) < Jatpunti (*Puntius sophore*) < Goldachingri (*Macrobrachium rosenbergii*) < Foli (*Notopterus notopterus*) < Boal (*Wallago attu*) < Gulsha (*Mystus cavasius*) < Bele (*Glossogobius giuris*) < Tengra (*Mystus vittatus*) < Bajari-tengra (*Mystus tengra*) < Chewa (*Pseudapocrypter elongates*) < Meni (*Nandus nandus*) < Borobaim (*Mastacembelus armatus*) < Poa (*Otolithoides pama*) < Bacha (*Eutropiichthys vacha*). The other fishes that analysed in one or two seasons showed the chronology, Kaikka (*Xenentodon cancila*) < Khalisha (*Trychogaster fasciata*) <*Gutum (Lepidocephalus guntea*) < Magur (*Clarius batrachus*) < Tarabaim (*Macrognatus aculiatus*) < Shole (*Channa striata*) < Gojar (*Channa marulius*). In each season, the variation of DDT contents varied with fish species, when the concentrations in a fish varied between seasons. This might be attributed to the combine influence of the trophic position, feeding habits, lipid contents, physiological activities (metabolism rate, excretion rate and maturation stage etc.) of fishes and meteorological parameters i.e. temperature, humidity and rainfall to the accumulation of DDTs in fish tissues. Overall, the rank orders of average DDTs of different fish and prawn species of different seasons were carnivore $>$ omnivore $>$ herbivore and summer $>$ autumn $>$ rainy-season $>$ winter. From the present study, it could be said that lower residues may be found in herbivores, lean and plant based omnivores and lean and zooplankton based carnivores; medium or higher residues may be found in fatty and animal based omnivores, lean and lower carnivores while much higher residues may be found in the fishes with bottom feeder carnivores, predators and fatty top carnivores. Analysis of total DDTs residues by one way ANOVA with LSD and Tukey HSD tests showed that significant differences in total DDTs between herbivore, omnivore and carnivore ($p < 0.05$). Highly significant differences between seasons ($p < 0.001$) were also observed. Pearson correlations analysis showed the positive relationships between DDE and DDD with lipid contents ($p < 0.05$). Besides, highly significant positive relationships between total DDTs with temperature with $(p < 0.001)$ and negative relationship between humidity with total DDTs ($p < 0.01$) were observed. DDD (55.54%) was the major contributor to fish and prawn samples in rainy-season and followed by 4,4́ -DDT (16.10%), DDE (14.35%) and 2,4́-DDT(14.00%). In autumn season, the major contributor was DDE (53.11%) and followed by the DDD (31.79%), 2,4'-DDT(8.52%) and 4,4-DDT(6.58%). In winter the major contributor was DDE (68.80%) and followed by DDD (18.23%), 2,4-DDT (7.63%) and 4,4-DDT $(5.33%)$. In summer, the major contributor was $4,4$ -DDT $(40.23%)$ and followed by 2,4́ -DDT(27.59%), DDE (18.60%) and DDD (13.58%). Compositional distribution of DDTs and the ratios of (DDE+DDD)/DDTs contributing to the values indicating both recent and past use of DDT in the region. The concentrations of total DDTs in all the samples were within the permissible Maximum Residue Level (MRL) i.e. for human consumption recommended by FAO-WHO. But 20.83% species of rainy-season, 68.18% of autumn, 13.04% in winter and 100% in summer exceeded the maximum admissible limit recommended by European Union. However, 4.55% species of autumn and 36.36% of summer were above concentrations associated with reproductive toxicity in several species of fish. To assess human health risks, Health risk Indexes (HI) of fish and prawn samples were calculated. HIs < 1 in all fish and prawn samples indicating that the fishes are safe to consume but the daily consumption of the fishes together with other contamination in food may cause human health hazard. As DDT is a long persistent and bioaccumulative substance in the environment, intake of significant amount of these poisonous elements with our diet is a matter of great health concern.

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1. INTRODUCTION

1.1. Background

Organohalogen compounds such as organochlorine pesticides (OCPs) represents a key group of persistent organic pollutants (POPs). One of the most hazardous and ubiquitous OCP is Dichlorodiphenyltrichloroethane (DDT). DDT and its metabolites are of great concern to the environmental scientist for several decades, due to their long-range transport, persistence, bioaccumulation, biomagnification, toxicity and adverse effects on environment and human health including reproduction and birth defects (Edwards 1987), endocrine disruptions, immune system dysfunction and cancer (Adeyemi *et al.* 2008). DDT is listed on Stockholm convention as a Persistent Organic Pollutant (POP) among 12 POPs where proposed the actions for elimination and restriction their fresh release in environments (UNEP 2001). Considering the adverse effects, many countries have restricted or banned their use (Wania and Mackay 1996, Sabljic 2001). In Bangladesh, DDT and its metabolites have been used extensively for agriculture as well as for mosquito control programs for long time. DDT was completely banned in the country in 1993 when their applications were officially phased out but it"s use remains illegally (Matin *et al.* 1998).

Bangladesh is an agro-based riverine country enriched with vast fisheries resources. Fish is an essential and mostly preferred food, other than rice in the country. It accounts for the largest share of per capita food expenditures after rice (Minten *et al.* 2010) and is by far the most frequently consumed animal-source food, providing approximately 60% of animal protein in the diet as well as other essential nutrients (Belton *et al.* 2011, Roos *et al.* 2007). Fisheries sector plays an important role in the national economy and nutrition as it contributes about 2.73% of the total export earning, 3.69% to GDP and 22.60% to agricultural sector (BES 2014). In case of inland fisheries production, Bangladesh is ranked fourth largest in the world (NFW 2015). Although rapid population growth and one of the world"s highest population densities (Talukder 2005), the country has attained close self-sufficiency in rice and vegetables. As a result, pesticide use in general is increasing for higher food

production to meet the demand of growing population. According to statistical data of the Government of Bangladesh, pesticides consumption increased from 7350 metric tons in 1992 (Rahman 2005, Rahman and Thapa 1999) to 16,200 metric tons in 2001, more than doubling in the past decade. (Rasul and Thapa 2003, Hossain 1988). The agricultural land lead to the contamination of aquatic environments by Persistent organic pollutants (POPs) mainly through the rain water runs off. Specially, fishes uptake pollutants through gills, skin and food intake which then transfer the contaminants to humans through consumption of these organisms (Zhou *et al.* 2008). As fishes of the country contribute great role in national nutrition and also exported in many countries of the world, detail research studies are needed to assess the occurrence and toxic effects of POPs of fishes while very few research have been conducted to such extent which give the true picture of contamination levels in the country as well as the region.

The Meghna River is one of the most important rivers in [Bangladesh,](https://en.wikipedia.org/wiki/Bangladesh) one of the three that forms the [Ganges Delta,](https://en.wikipedia.org/wiki/Ganges_Delta) the largest on earth fanning out to the Bay of Bengal. In the Sonargaon Upazila of Narayanganj District, the Meghna River plays a vital role in the economy of the district as they facilitate fisheries resources, irrigation, drainage, and water supply of the area. On the bank of the Meghna River at Sonargaon Upazila different types of rice, jute, vegetables, banana, betel leaf, sugar cane etc. are cultivated throughout the year (NDS 2013). Pesticides are used improperly due to lack of appropriate knowledge about their applications and untoward effects. Moreover several mills, factories and industries are present on the bank of the river in the upazila. It is obvious that the wash out of pesticide residues by rain water and industrial effluents discharge contaminate the aquatic environments of the rivers and lakes. These contaminants can be potentially bioaccumulated in the fatty tissues of fish and biomagnified from lower trophic level to the higher trophic levels through food chain.

This research aims to determine amount of organochlorine pesticides (OCPs) such as DDT and its metabolites residues in fishes of different trophic levels of the Meghna River during different seasons.

1.2. Persistent Organic Pollutants (POPs)

Persistent organic pollutants are organic chemical substances, possess some physical and chemical properties in a particular combination therefore once released into the environment, they may

- remain intact for exceptionally long periods of time.
- transport over large distance and become widely distributed throughout the environment involving air, soil and water.
- accumulate in the fatty tissue of living organisms including human, biomagnified through the food chain that found at higher concentrations at higher trophic levels.
- pose a risk of causing adverse effects to the environment and human health.

POPs include pesticides such as DDT, industrial chemicals such as polychlorinated biphenyls (PCB) and unintentionally generated chemicals such as polychlorinated dibenzofurans (PCDF) and polychlorinated dibenzo-*p*-dioxins (PCDD). Over the past several decades, POPs are released to the environment through human activities. Due to the nature of semi-volatile, long half lives and long range transport, POPs are now extensively distributed throughout the world including some of the most remote areas (Daly *et al.* 2007, Fernandez and Grimalt 2003, Lohmann *et al.* 200, Ondarza *et al.* 2011) and the areas where POPs have never been used (eg. Polar region). POPs are accumulated in the fatty tissue of living animals and human beings therefore the concentrations of POPs residues in fatty tissue can become magnified by up to much higher (About 70 000 times) than the initial levels. Because of it's biomagnifications property through the food chain, concentrations of POPs tend to increase so that animals at the top of the food chain such as fish, predatory birds, mammals, and humans tend to have the greatest concentrations of these chemicals, and therefore are also at the highest risk from acute and chronic toxic effects (Fernandez and Grimalt 2003).

POPs are highly toxic which enter in to environment through human activities but mainly unintentionally (Kennish 1997, Breivik *et al.* 2004).They negatively affect humans, plant and animal species and natural ecosystems both in close proximity and

at significant distances away from the original source of discharge. Exposure to POPs in humans, either acute or chronic, can be associated with a wide range of adverse health effects, including illness and even death. The adverse effects of POPs can include hypersensitivity, allergies, disruption of the immune system and endocrine, damage to the central and peripheral nervous systems, reproductive disorders and carcinogenicity (El Nemr *et al.* 2011, Amodio *et al.* 2012, El Nemr 2013). POPs cause endocrine disruption by altering the hormonal system and damage the reproductive and immune systems of exposed individuals as well as their offspring. Moreover the higher levels of POPs residues in human blood serum can be associated to diabetes (Lee *et al.* 2006).

According to the U.S Environmental Protection Agency (EPA) there are links between POPs exposure and the increased frequency of diseases and abnormalities in wildlife species, including certain kinds of fish, birds, and mammals. The negative effects of pesticides in the marine and coastal environments include changes in reef community structure, such as decreases in live coral cover and increases in algae and sponges and damage to sea grass beds and other aquatic vegetation from herbicides (EPA 2012).

Considering the adverse effects of POPs, great concern has arisen about the occurrence of the POPs in environment for the last decade. Scientists and policymakers around the world became concerned about these harmful compounds and decided to make a convention in a meeting in Sweden on May 23, 2001 known as the "Stockholm Convention". In May 2001, the [United Nations Environment](https://en.wikipedia.org/wiki/United_Nations_Environment_Programme) [Programme](https://en.wikipedia.org/wiki/United_Nations_Environment_Programme) (UNEP) through the Stockholm convention investigated POPs, initially the Convention recognized only twelve POPs i.e. aldrin, chlordane, dieldrin, DDT, endrin, heptachlor, mirex, toxaphene, hexachlorobenzene, PCBs, PCDDs and PCDFs as the primary problem compounds, and are commonly known as the "Dirty dozen". At present, the number of POPs has increased to 19. The aim of the convention is to protect human health and the environment by phasing out of these hazardous pollutants from the earth.

1.2.1. Classification of persistent organic pollutants

POPs are classified into two important subgroups including both the polycyclic aromatic hydrocarbons and some halogenated hydrocarbons. The halogenated group includes several organochlorines which are generally the long term persistent of all the halogenated hydrocarbons.

1.2.1.1. Organohalogen compounds (OHCs)

Organic halogen compounds are a large class of natural and synthetic chemicals that contain one or more halogens (fluorine, chlorine, bromine, or iodine) combined with carbon and other elements. Organohalogens are widely used in industry and society. Most of the organohalogen compounds in use recently are synthetic in origin. However, huge amount of synthetic halogen compounds are used as pesticides, aerosol propellants, cleaning solvents, anaesthetics, polymers, refrigerants and so on. The wisdom of this massive use of materials that are foreign to our natural environment gradually is being reevaluated as the long-term detrimental effects of many of these chemicals become known. For example, many of the chlorinated hydrocarbons such as DDT, Chlordane, and Lindane, which have been used very widely as insecticides, now are at least partially banned because of concern for their long-term effects on non-target species, including man.

1.2.1.2. Organochlorine pesticide residues (OCPs)

The chemical compounds used to control pests and diseases of plants, to eradicate weeds, to kill pests and microorganisms that spoil agricultural products, materials and articles, and to control parasites and vectors of dangerous diseases of human and animals are called pesticides (Grugdyev *et al.* 1983). The term pesticide includes insecticide, herbicide, acaricide, ovicide, larvidice, molluscicide, nematicide, rodenticide, fungicide, bactericide, etc. Pesticides can be classified according to chemical class such as organochlorine, carbamate, organophosphorus, chlorophenoxy compounds.

Organochlorine pesticides are chemically produced insecticides composed primarily of carbon, hydrogen and chlorine which include [DDT,](http://www.coastalwiki.org/wiki/DDT) dieldrin, heptachlor, chlordane, [endosulfan](http://www.coastalwiki.org/wiki/Endosulfan) and [dicofol.](http://www.coastalwiki.org/wiki/Dicofol) Their carbon-chlorine bond is very stable towards hydrolysis and, the greater the number of chlorine substitutions, the greater the resistance to biological and photolytic degradation. By virtue of their high degree of halogenations, they have very low water solubility and high lipid solubility, leading to their propensity to pass readily through the phospholipid structure of biological membranes and accumulate in fat deposits. Chlorinated pesticides, with molecular weights 28 greater than 236 g/mol, have the ability to accumulate in biological tissues and to concentrate in organisms that occupy positions in the upper trophic levels. The general characteristics of OCPs; high chemical stability, low volatility, high lipid solubility, slow biotransformation and degradation, make them persistent, bioconcentrate and biomagnify through food chain (UNEP 2001). Finally these persistent compounds cause several environmental and human health hazards. Considering these adverse effects, the worldwide environmental scientists and general public pay their attention to these pollutants for the last decade (Law *et al.* 2003, Jiang *et al.* 2009).

OCPs were heavily used from the mid-1940s to the mid-1980s. The persistence of OCPs, their bioaccumulation tendency and global contamination resulted in their ban and restriction in most countries. Banned OCPs include DDT, aldrin, dieldrin, toxaphene, chlordane and heptachlor. Although the production and use of OCPs have been banned in most of the country but they are still used illegally in some developing countries where there are evidence of their continued usage and presence in ecosystems (Darko and Acquaah, 2007). In some countries like Africa, Malaysia, Chine and India, they are still used in many sectors including agricultures, industry and public health (FAO 1999, Karlsson *et al.* 1997). However, the extensive use of OCPs for the past several decades and recent illegal use in agriculture and public health sectors has resulted an accumulation of these toxic residues in various environmental compartments such as soil, sediment, water, fish, birds, wildlife, vegetables and human body (Nishina *et al.* 2010, Sharma *et al.* 2013, Wang *et al.* 2013, Osafo and Frempong 1998, Ntow 2001).

Generally pesticides are used in terrestrial ecosystems which then flushed away through rain water to rivers and end up in marine ecosystems. Moreover the deposition of organochlorine pesticides in the environment occurs through several input mechanisms of which include urban runoff, municipal, sewage, industrial waste, outflow from agricultural areas, chemical spills, and atmospheric deposition (Vallack *et al.* 1998). After enter into the environment, they pass through food chain and enter into organisms through their food. In general, the principal route of human exposure to OCPs is dietary intake, accountin for >90% as compared to dermal and inhalation exposure pathways (Mansour *et al.* 2009)

As mentioned before, despite their restrictions, these compounds are still detected in the environment and in tissue samples such as blood, adipose tissue and breast milk of humans. But recently, body burdens have declined since these organochlorines were banned, yet virtually the entire population still carries detectable levels of the toxic chemicals. Chronic exposure to low levels of OCPs can cause a wide range of serious harmful effects in animals and humans. However, a number of studies have demonstrated that pose serious health hazards to human (Adeyemi *et al*. 2008, Nakata *et al.* 2002, Sverdrup *et al*. 2002, Jayashree and Vasudevan 2007). The chlorinated organic pesticides can pass through the mother placenta to the unborn child (Nakata *et al*. 2002). They cause many harmful effects such as abnormal development of the immune system, birth defects and fetal death (Ayejuyo *et al.* 2008). Thus organochlorine pesticides are considered as one of the main environmental and human health problems in the world (Darko and Acquaah 2007, Doong *et al*. 2002).

1.3. Dichlorodiphenyltrichloroethane (**DDT)**

DDT is an organochlorine insecticide. As an important group of POPs, DDT is mainly characterized by highly toxicity, long range transport, long life span and lipophilic properties. Because of it"s lipophilic nature, it can be accumulated in the fat body of living organisms which then biomagnified to the top consumers than the lowers. Finally this hazardous chemical causes serious health effects in human and threat to environment.

1.3.1. Physical properties

DDT is a white crystalline powder with minimal odor. It is nearly insoluble in water but soluble in fats, oils and most organic solvents. DDT does not occur naturally, but is produced by the reaction of chloral (CCl₃CHO) with chlorobenzene (C_6H_5Cl) in the presence of sulfuric acid, which acts as a catalyst. The Chemical formula for DDT is $C_{14}H_9C_{15}$. It's melting point is 108.5 °C; boiling point is 185 at 0.05 mm Hg and degradation point is above 250 $^{\circ}$ C. Its structure is similar to the pesticides dicofol and methoxychlor.

1.3.2. Isomers of DDT

Commercial DDT is a mixture of some related compounds. The components include the 4,4′-DDT isomer (77%), 2,4′-DDT (15%), dichlorodiphenylethane (DDE) and dichlorodiphenyldichloroethane (DDD). In the environment and in the body, dichlorodiphenyltrichloroethane (DDT) breaks down into dichlorodiphenylethane (DDE) and dichlorodiphenyldichloroethane (DDD) over time. DDE and DDD are the major metabolites of DDT. The total DDT in a sample refers to the sum of all DDT congeners (4,4′-DDT*,* 2,4′-DDT, DDE and DDD) (WHO, 1973 and Nahar *et al*. 2008). From this context, the term Σ DDTs (Σ is used to mean sum of) will be used. In some cases, the DDT will be used to refer to the collection of all forms of DDT, DDE and DDD.

Table 1.1. Physical properties of DDT and its Metabolites (Beyer 2000)

1.3.3. History of DDT

DDT is one of the well known synthetic pesticides. In 1874, it was first synthesized in the laboratory by a German chemist named Othmar Zeidler. Then Paul Muller discovered its insecticidal properties in 1939 (Mischke *et al*. 1985). DDT was first manufactured in 1943 (Pretty and Hine 2005). During Second World War, it was used by the military and civilians to control the spread of malaria and typhus by mosquitoes and lice respectively.

After the war, during 1945, DDT was used as agricultural pesticide and also to control insect-borne human diseases (EPA 2012). From that period, DDT was used extensively for around thirty years. Increasing environmental and human health hazards causes by this persistent and toxic chemical made the scientist concern about it (Carson 1962, Longnecker *et al.* 2005). Therefore, during 1959, the use of DDT as agricultural pesticides and also vector control insecticide was declining and since 1970s DDT was banned in most of the developed countries though it is still being used in some underdeveloped countries for disease vector control (Biscoe *et al.* 2004). Even so, in many developing countries still DDT is used illegally and also some countries produce DDT for agriculture purpose and malaria vector combat (Biscoe *et al.* 2004).

1.4. DDT, DDE, and DDD in to the environment

1.4.1. How DDT enters into the environment

DDT is manufactured chemically which is not known to occur naturally in the environment (WHO 1979). Historically, DDT was released to the environment during its extensive use as a pesticide in agriculture and vector control application. Because of its persistence and long half life, most DDT in the environment is a result of past use but still DDT enters into the environment due to its current use in different areas of the world. After sprayed on the crop as pesticide or use to vector control, this pollutants release in the environment and persist in the air, water, soil and organisms which then cyclically move through food chain, biomagnified and finally cause hazard to biota.

1.4.1.1. Air pollution

DDT may be released in land or in atmosphere. In case of land region, from soil and water, some parts of released DDT evaporate to the air. When in the air or atmosphere, about 50% of DDT will be adsorbed to particulate matter and remaining will exist in vapor phase (Bidleman 1988). During the journey in soil, water and air, some amounts of DDT may be changed to DDD and DDE. However DDT, DDE, and DDD in the air will then be deposited on land or surface water again. This cycle of evaporation and deposition may be repeated many times which would help to

transport them to long distances in the atmosphere. Therefore, these chemicals have been found in bogs, snow, and animals in the Arctic and Antarctic regions, far from where they were ever used.

1.4.1.2. Soil pollution

After application in terrestrial region, some DDT may have entered the soil from waste sites. Even though some parts of soil"DDT may be evaporated to air, most of them stick strongly to soil particles, and therefore generally remain in the surface layers of soil. Most DDT breaks down slowly into DDE and DDD, generally by the action of microorganisms.

DDT, DDE, and DDD last in the soil for a very long time, potentially for hundreds of years. The length of time that DDT will last in soil depends on many factors including temperature, type of soil, and whether the soil is wet. DDT disappears faster when the soil is flooded or wet than when it is dry. DDT disappears faster when it initially enters the soil. DDT lasts for a much shorter time in the tropics where the chemical evaporates faster and where microorganisms degrade it faster. In tropical areas, total DDTs (ΣDDTs) may disappear in much less than a year. In temperate areas, half of the ΣDDTs initially present usually disappear in about 5 years. However, in some cases, half of the ΣDDTs initially present will remain for 20, 30, or more years.

Some soil particles with attached DDT, DDE, or DDD may get into adjacent rivers and lakes in runoff. Only a very small amount, if any, will seep into the ground and get into groundwater. DDT in soil can also be absorbed by some plants and by the animals or people who eat those crops.

1.4.1.3. Water pollution

DDT enters into the aquatic habitat through surface run off from contaminated lands and also deposition from air. After entering in the water body, it will bind to particles which then settle down and be deposited in the bottom sediment. From the water column, DDT is taken up by phytoplankton, zooplanktons, small organisms, fishes, molluscs and other aquatic organisms. It accumulates to high levels in fish and marine mammals (such as seals and whales) through repeated feeding of other organisms. In those animals, level of DDTs residues may be reached many thousands of times higher than in water.

1.4.1.4. Fish pollution

Aquatic ecosystems are the ultimate reservoirs of many contaminants including DDT, from where contaminants can easily be accumulated to aquatic organisms due to their (contaminants) persistent and lipophiclic properties (Ondarza *et al.* 2012). Particularly, fishes are able to uptake contaminants directly from water and or through their food chain. Therefore, fish can bioaccumulate DDT by two main routes in their natural aquatic habitat; from water via gill and body surfaces and from food intake (Fisk *et al.* 2001, Blocksom *et al.* 2010, Ondarza *et al.* 2011, Poma *et al.* 2014). However fishes with organochlorine residues can reflect the pollution status of the surrounding environments as the metabolisms rate of DDT in fish is moderate (Guo *et al.* 2008, b). Additionally, fish from the same site may occupy different trophic positions in their food chain, reflecting in different degrees of contaminant biomagnification (Hoekstra *et al.* 2003). After uptake contaminants, fishes eventually transfer them to piscivorous birds, wild animals and human through consumption of these organisms (Zhou *et al.* 2008).

1.4.1.5. Human in the food chain

People may be exposed to DDT, DDE, and DDD through several ways. During the handling or application, DDT could be entered into the body through inhalation or absorption through the skin. While the common route of exposure is through the food intake containing small amounts of these compounds. Particularly fatty foods such as fish, meat, milk, poultry and dairy products contain DDTs residues. In case of vegetables, leafy vegetables contain more DDT than other vegetables, possibly because DDT in the air is deposited on the leaves.

After enter into the body, DDT can break down to DDE or DDD. Because of lipophilicity these contaminants are stored mostly in the fat body, especially DDE. Some of these stored amounts leave the body very slowly, mostly in urine or by breast

milk to the newborn. The amount of DDT in the body decreases with decrease exposure. But DDT levels may be increased with continuous exposure and may cause health hazard.

Mother can transfer DDT to her unborn baby through placenta or to the newborn through breast milk. However, in most cases, the benefits of breastfeeding outweigh any risks from exposure to DDT in mother"s milk (Bouwman *et al.* 1990).

1.5. Toxicology of DDT and its metabolites

DDT was first synthesized in 1874 but its insecticidal properties were not discovered until 1939 (Smith 1991), and large scale industrial production started in 1943.

DDT is given credit for having helped 1 billion people live free from malaria, thus saving millions of lives. In 1973, after 30 years of worldwide use of DDT, a World Health Organization (WHO) report concluded that the benefits derived from use of this pesticide were far greater than its possible risks (WHO 1973). After 25 additional years, the benefits of DDT can be confirmed, but its stability, ubiquitous presence, and persistence in the environment, its accumulation in adipose tissues, and its estrogenic properties raise concern about its possible long-term adverse effects. In addition to a possible carcinogenic effect, DDT has been reported to affect neurobehavioral functions and to be associated with premature births (VanWendel *et al*. 2001, Longnecker *et al*. 2001). No living organism may be considered DDT free. DDT is stored in all tissues, but the highest concentration occurs in fat. It has been calculated that it would take between 10 and 20 years for DDT to disappear from an individual if exposure would totally cease, but that DDE would possibly persist throughout the life span (Smith 1991).

1.5.1. Human health hazard

1.5.1.1. Short term health hazard

If anybody takes large amounts of DDT over a short time by eating food or inhalation then he would be experienced sweating, headache, nausea, vomiting, and dizziness. These could finally affect nervous system, reproductive system and also adrenal glands.
1.5.1.2. Long term health hazard

People exposed for a long time to small amounts of DDT (less than 20 mg per day), such as people who worked in factories where DDT was made, had some minor changes in the levels of liver enzymes in the blood. A study in humans showed that increasing concentrations of DDE in human breast milk were associated with reductions in the duration of lactation. An additional study in humans found that as the DDE levels in the blood of pregnant women increased, the chances of having a pre-term baby also increased.

Animal studies show that long-term exposure to moderate amounts of DDT (20-50 mg kg-1 of body weight every day) may affect the liver. Studies in animals have shown that oral exposure to DDT can cause liver cancer. The International Agency for Research on Cancer (IARC), the Department of Health and Human Services and also the Environmental Protection Agency (EPA) together has determined that DDT, DDE, and DDD are possibly carcinogenic to humans.

1.5.1.3. Experimental evidence of carcinogenicity of DDT and its metabolites

The carcinogenicity of DDT and its metabolites has been studied in a number of laboratories in animals including non human primates. DDT induced tumors in mice after treatment for a limited period of time. When treatment was continued for 15 or 30 weeks and mice were killed at various intervals, similar proportions of mice bearing liver tumors were observed 65, 95, and 130 weeks after the beginning of the experiment. More mice had large liver tumors at 95 and 120 weeks than at 65 weeks. These results indicate that although no new tumors were induced after cessation of exposure to DDT, the persistence of DDT-induced hematomas did not depend on the continuous administration of DDT because the tumors that have already appeared continue to grow (Tomatis *et al.* 1974, Tomatis 1975). After treatment of mice with DDT for six generations at four doses, no increase in the incidence of liver tumors was observed from generation to generation as might have been expected in the presence of a genotoxic carcinogen (Turusov *et al.* 1973). DDT also increased incidences of lung tumors and lymphomas in mice, incidences of liver tumors in rats, and incidences of adrenal adenomas in hamsters (IARC 1991). Long-term oral

administration of DDT to nonhuman primates was reported to result in hepatic toxicity, and a few malignant and benign tumors at various sites were also found at an incidence that was of borderline statistical significance (Takayama *et al.* 1999). In the absence of tumors in the controls, the few tumors observed may be considered to be biologically significant, thus confirming the carcinogenic effect observed in rodents (Tomatis 2000).

A number of reports have indicated that organochlorine insecticides, including DDT and its metabolites may act as endocrine disruptors (Keith *et al.* 1997, Longnecker *et al.* 1997). Because a considerable proportion of all cancers in women are hormonally mediated, the possibility that xenoestrogenic substances, such as organochlorine insecticides, contribute to an increased cancer risk is particularly alarming (Wolf *et al.* 1996). Early reports showed higher concentrations of DDT and DDE in fat tissue of individuals with mammary cancer (Wasserman *et al.* 1974) and an association between DDE blood levels and mammary cancer (Wolf *et al.* 1993, Dewailly *et al.* 1994, Krieger *et al.* 1994).

1.5.3. Effects on wildlife

Exposure to DDT through the environment and/or food chain, may alter the immune system of birds and mammals; induce thyroid dysfunction in shellfish and mammals; decrease fertility rates in some wild animals (e.g. mink) and may also cause disruptions in the sex characteristics of individual animals, thereby altering the sex ratio of the population (Colborn and Smolen 1996). DDT leads to great deal of negative effects in the wildlife such as on eggshell thinning of herring gull, peregrines, falcons and hawks; led to raised level of endocrine disrupting chemicals in the alligator's tissue and a decrease of 50% of juvenile alligator numbers; resulted to the disease syndrome in grey and ringed seals in Baltic which caused to decline of seal population; DDT is also a neuro-developmental toxicant with a lot of evidences in changes of behavior and neurochemistry into adulthood of mice which exposure to DDT in stage of pre-natal and neonatal nervous system development. The endocrine disrupting effects of DDT may be one of their major impacts on wildlife (Thuy 2015).

1.5.4. Effects on fish species

DDT is highly toxic to fish species. Schoenthal. in 1963 reported that the mortality rates of rainbow trouts were increased by feeding aquatic insects treated with DDT. DDT also induce thyroid dysfunctions in fish and aquatic mammals (Colborn and Smolen 1996). Atlantic salmon exposed to DDT as eggs experienced impaired balance and delayed appearance of normal behavior patterns. DDT also effects on temperature selection in fish. Moreover DDT might lead to toxic effects in fish, causing endocrine disruption and altering biochemical, physiological, histological and morphological parameters (Norena-Barroso *et al.* 2004, Da Cuna *et al.* 2011, 2013, Hued *et al.* 2013).

1.6. Occurrence of DDT in different country

DDT have penetrated almost all the ecosystems and are now ubiquitous; this is evidenced by their detection in all environmental compartments and biota (Aono and Tatsukawa 1997, Norstrom *et al*. 1998, Muir *et al.* 2000). Considering this condition, environmental scientists have done several research works on the residue levels of DDT in water, soil, sediment, vegetables, fishes, other animals and also the fat containing food in different part of the world. The following tables would help to represent the world wide conditions of DDT levels in different compartments of the environment and also biota.

Table 1.2. DDTs in the water samples from different countries

Table 1.3. DDTs in the soil/sediment samples from different countries

*dw- dry weight

Table 1.4. DDTs in the fish samples from different countries

Table 1.5. DDTs in the human and different biota samples from different countries

*ww-wet weight, lw-lipid weight

1.7. Organochlorine pesticides (OCPs) in Bangladesh

1.7.1. Environmental legislations for OCPs in Bangladesh

According to the Pesticide Ordinance 1971, powers were given to the government in order to make pesticide rules. Pesticide Technical Advisory Committee made the pesticides rules in 1985 which gave the authority to the Director of Plant Protection Wing of Department of Agriculture Extension (DAE) for the registration of pesticides that started in 1986 in the country in exercise of clause 4 and clause 5 of the ordinance. The Pesticide Rules, 1985 provide the registration of manufacturing or importing of pesticides valid for three years (Rahman *et al.* 1995). Due to the this registration is valid for three years, but unfortunately the practice of assessment is not properly undertaken owing to (Gaston 1986). The implementation of the rule also face disappointing situation as huge amount of toxic and banned pesticides has smuggled from India. According to Parveen and Nakagoshi (2001), the main drawback of the above said regulation is in chapter VII section 33 sub section I (a) which actually provides the provision to state the manufacturer name, formulator name or repacker name in the label even in a case when the certain pesticide is not registered on his/her name. This could make the identification of a respective person very difficult. Taking the advantage of this weak point present in the regulations, the illegal business in the country is going on. The Bangladesh Government signed the Stockholm Convention on 23rd May, 2001. As a signatory, the Government was committed to prepare their National Implementation Plan for POPs, take action for creating awareness regarding consequences of POPs releases and ultimately their elimination from the environment. For National Implementation Plan, the concerned authorities were Department of Environment of Ministry of Environment (ESCO 2005).

1.7.2. DDT in Bangladesh

The use of pesticides, including organochlorine compounds, in Bangladesh started during the middle of the 1950s to promote crop production. A factory for production of DDT was built during the 1960s. Bangladesh had small pesticide use until 1970s, farmers were motivated to use pesticide free of cost up to 1974 and at reduced price till 1980 (Rahman 2000). According to statistics from the Government of Bangladesh, consumption of pesticides increased from from 2200 million tons in 1980–82 to 6500 million tons in 1992–94 (Rahman 2005, Rahman and Thapa 1999). According to Rahman *et al.* (1995), 2510 tons of pesticides were used in 1982 and 5150 tons in 1988 which increased to 8000 tons in 1994 and then 16,200 metric tons in 2001, more than doubling in the past decade Matin *et al*. (1998) reported that restricted use of OCPs was allowed by the Registration Authority of Bangladesh in 1990. They also reported that OCPs consumption for public health and malaria control program was low as compared to agricultural usage and 100 tons formulated OCPs were used in 1990 for malaria control programs. Because of the risk to the human health and environment, a number of organochlorine pesticides were banned in Bangladesh in 1993, including DDT. The factory producing DDT was closed down while in 1994, DDT was allowed by public health members for immediate control of plague. However, report says that DDT is still being illegally used in the country (Takada *et al.* 2003).

Table 1.6. DDTs in water, soil, vegetables, fish and human blood samples of Bangladesh

1.8. Objectives of the present study

Fishes are the most important and preferable food item not only in Bangladesh but also all over the world considering their high nutrient values. Yet at the same time, fishes may contain several contaminants such as organochlorine pollutants taken from the surrounding contaminated environments. Therefore the people should be informed of both the benefits and risk of fish consumption.

The developments of Fisheries sector of Bangladesh are now worldwide recognized. Fishes contribute important roles in the both national and abroad nutrition through high production and huge export in the foreign countries. Therefore it is necessary to have the detail and clear information about their nutrients and contamination profile. In our country, several research works have done on their nutrient evaluation while very few or not so detail works on the toxicological profile of fishes.

From this point of view, the present study investigated the organohalogen residues (specially DDT an its derivatives residues, the most toxic and hazardous persistent pollutants) of different fishes of different seasons. The objectives of the present study are as follows

- To analyze organohalogen pesticide residues from different fishes along the Meghna River in Sonargaon Upazila, Narayangang
- To analyze the seasonal variation of organohalogen residues in fishes from the Meghna River in Sonargaon Upazila
- To correlate organohalogen residues in fishes with selected meteorological parameters i.e. temperature, rainfall and humidity
- To analyze the bioaccumulation of organohalogen residues in fishes at different trophic levels

2. MATERIALS AND METHODS

2.1. General

Meghna is one of the most important river in Bangladesh, deteriorating rapidly due to pollution from the wash-out of chemical and fertilizer of agricultural land and the effluents of industries on the bank of the river. The present study aims at determining the organohalogen residues in fishes from Meghna River and also the biological accumulation through the food chain. To fulfil the objectives of this research, the following methodologies were adopted.

2.2. Materials

2.2.1. Chemicals and solvents

Anhydrous magnesium sulphate were purched from Junsei Chemical Co. Ltd., Japan; anhydrous sodium sulphate, sodium chloride (analytical grade) and concentrated H2SO⁴ (Merck, Germany) were used for this analyses. Acetone, *n*-hexane and ethyl acetate were purchased from SK Chemical Co. Ltd, Republic of Korea.

2.2.2. Standard compounds

The standards of 2,4*´*-DDT & 4,4*´*-DDT (99% purity), 4,4*´*-DDE (99% purity), 4,4*´*- DDD (99% purity), purchased from Dr. Ehrenstorfer, Germany were used for analysis.

2.2.3. Equipment

Calibrated balance, volumetric flasks and pipettes (calibrated by BSTI) were used for the analysis. All required glass apparatus were cleaned with water using detergent, rinsed about six times with water, then twice with distilled water and finally with redistilled acetone. All glassware was baked at 300ºC overnight and stored by covering with aluminum foil prior to use.

2.2.4. Instruments

2.2.4.1. Homogenizer

Fish samples were homogenized by normal kitchen blender (Panasonic, China).

2.2.4.2. Centrifuge machine

The samples were centrifuged by Hanil Science Industrial Co. Ltd., Model-Combi 514 R or by Heraeus Sepatech (Labofuge A).

2.2.4.3. Gas Chromatograph

A Shimadzu-2010 Gas chromatograph with electron capture detector and auto injector (ECD) was used for the determination of pesticide residues in the samples. A HP-5MS quartz capillary column (30 m \times 0.25 mm i.d. and 0.25 µm film thickness) from Agilent, USA was used to carry out the separation. Nitrogen was used as the carrier and make up gas. All injections were made in split-less split mode and injection volume was 1 μ L.

 Fig. 2.1. Shimadzu GC-2010

2.2.5. Methods

2.2.5.1. Activation of chemicals

Sodium sulphate and Magnesium sulphate were activated by heating at 300 °C for 8 h in a furnace (GSM 11/8 Hope valley, S336RB, England) then kept out from furnace

and were allowed to cool at room temperature. After that the Sodium sulphate and Magnesium sulphate kept in a desicator.

2.2.5.2. Saturation of sulphuric acid

Concentrated sulphuric acid (80 mL, 98%) was taken in a reagent bottle and 20 mL of n-hexane was added to the acid. It was shacked about one minute and kept for 5-6 minutes to separate the two phases. The lower sulphuric acid phase, saturated with nhexane was collected and stored in amber colored bottle.

2.2.5.3. Evaporation

All the evaporations were carried out under reduced pressure using rotary evaporator (Büchi Rotavapor R-114, Germany or Heidolph, Germany) at water bath temperature not exceeding 40°C. Before evaporation of each sample extract, the rotary evaporator was washed for three to four times through evaporation of acetone.

2.3. Preparation of standard solutions

2.3.1. Preparation of primary standard solutions

The known amount of the certified pesticide standars was dissolved in a definite volume of n-hexane and the concentration of the standard was calculated. Primary stock solutions (100 µg/g) of 2,4´-DDT, 4,4*´*-DDT, 4,4*´*-DDE and 4,4*´*-DDD (99% purity) were prepared separately by dissolving 10 mg of each analyte in 100 mL nhexane. The prepared solutions in 100 mL were labelled indicating name of each of the standard, concentration and the date of preparation. The meniscuses of the solutions were marked with permanent black ink and stored in the freezer $(-20 \degree C)$ away from the sample storing area until further use.

2.3.2. Preparation of middle and working standard solutions

The primary standard solutions were taken out of the freezer to reach room temperature and checked the meniscus of the layer. If the meniscus of the layer was below the mark then adjusted with n-hexane and vortexed for one minite. Then a definite amount of solution was withdrawn and put a new mark in the stock solution after withdrawing. The withdrawn solution was diluted with the solvent appropriate

for making 20 µg/L secondary standard. Then 5 µg/L working standards were prepared from the secondary standard by diluting with solvent following same way as mentioned above. The working standard solution (5µg/L) was serially diluted to 2 μ g/L, 1 μ g/L in the same procedure.

These solutions were labelled indicating name of the standard, concentration and date of preparation. The meniscuses of the solutions were marked with permanent ink and stored in a freezer (-20°C) away from the pesticide residue laboratory.

 Fig. 2.2. Pesticide standard of DDT

2.4. Calibration curves

Working standards solutions of DDT and its metabolites were analyzed with GC-ECD and peak areas of corresponding solutions were listed. The calibration curves of each of the standard was prepared by plotting area vs concentration using Microsoft Excel-2010 software and r^2 were found to be 0.9960, 0.9950, 0.9840 and 0.9960 for DDE, DDD, 2,4´-DDT and 4,4´-DDT respectively.

Fig. 2.3. Calibration curves of DDE, DDD, 2,4´**-DDT and 4,4**´**-DDT**

2.5. Selectivity, Sensitivity and Linearity

Selectivity (or specificity) was assessed by analyzing standard mixture of pesticides, blank matrices and blank matrices spiked with the mixture of pesticides simultaneously and by checking their retention times. Sensitivity of the instruments was assessed by determining limits of detection (LODs) and limits of quantification (LOQs) for each pesticide in each matrix. Linearity was evaluated by constructing calibration curves for each pesticide by injecting standard mixture to GC at 5-8 different concentration levels covering the expected range of pesticides that might be present in the samples.

To determine the LOD, working standard solutions were serially diluted to get desired concentration. The diluted standard solutions were injected one by one until the peak heights of the standards were same to the noise level. The limit of detection (LOD) of the test compounds was determined using a signal-to-noise ratio of 3 with reference to the background noise obtained for the blank sample, whereas the limits of quantification (LOQ) were determined with a signal-to-noise ratio of 10.

2.6. Identification and Quantification by GC

The reference standard solutions were injected into the GC-ECD and under the same condition cleaned extract of samples were also injected. Comparing the retention times (retention time of standard and unknown supposed to be same under the identical analytical conditions) of the different peaks of the sample with the retention times of the standard compounds, corresponding residues present in the samples were identified. Quantitative determination was carried out by comparing the peak area of the each DDTs in the sample extract with that of the peak area of the respective DDTs in the external standard solution. For quantification, concentration of the corresponding analytes was found from standard calibration curve taking into consideration that the peak area was in the midpoint of the curves.

2.7. Blank experiment

For recovery experiments, control samples were used, which were previously confirmed that they had no pesticide. The control samples were spiked with known amount of DDTs were extracted follows three replicates followed by respective extraction and clean-up procedure to determine the matrix effect under analysis method. Reagent blank was done following the same extraction procedure and cleaned up method, using only solvent and reagents (in the absence of sample) to make the analysis rational.

2.8. Recovery Experiment

The recovery experiments were conducted on uncontaminated control samples by spiking the sample at 3 replicates in three concentration levels. The spiked samples were permitted to equilibrate for 3 h before extraction, in order to allow the pesticide penetrates the matrix. Then the spiked samples were subsequently processed by following the respective extraction and clean-up procedures. The recovery of the each analyte was calculated according to:

$$
R = \frac{A_m \times C_{st}}{A_{st} \times C_m} \times \frac{100}{M_{st}}
$$

Where *R* is the recovery $(\%)$, A_m is the peak area of the analyte in the matrix, A_{st} is the peak area of the analyte in the standard, *C^m* is the concentration of the analyte in the matrix, C_{sr} is the concentration of the analyte in the standard, and M_{st} is the mass of the analyte in the standard.

2.9. Method Validation

The extraction efficiency of the analytical procedure was evaluated via recovery experiments. Validation of the method was performed in terms of recovery studies before analysis of field samples.

2.10. Sample collection and preparation

2.10.1. Selection of sampling site

Fish samples were collected from the Meghna river at Boidyer Bazar of Sonargaon Upazila of Narayanganj District. Boidyer Bazar is the fish landing centre on the bank of the Meghna River at Sonargaon Upazila where around one hundred fishermen

bring the fishes caught from nearby the Narayanganj District of Meghna Rriver. The fishes brought there to sell to wholesaler at two times per day, fishes caught at night, brought at 5 a.m.- 6 a.m. and caught during morning, brought at 12 a.m. -1 p.m.

Meghna River is one of the major river in Bangladesh, specially famous for its great estuary that discharges the flows of the Ganges-Padma, the Brahmaputra-Jamuna and the Meghna itself. The Meghna has two distinct parts. The Upper Meghna from Kuliarchar to Shatnol is a comparatively small river. The Lower Meghna below Shatnol is one of the largest rivers in the world because of its wide estuary mouth. A larger number of settlements, towns, ports and industries have sprung up on both the banks of the Meghna. Narayangang, Narsingdi, Chandpur, Barisal and Bhola are the district towns that stand on the banks of the Meghna. Kuliarchar, Bhairab Bazar, Chandpur (Puran Bazar), Ramdaspur, Kalupur and Daulatkhan are important riverports and business centres. The Ashuganj thermal power plant and the Fenchuganj fertiliser factory are located on the banks of this river.

2.10.2. Selection of fish samples, period and way of sampling

Fish samples of different trophic levels; herbivore, omnivore and carnivore were collected to determine the bioaccumulation through food chain. In case of each fish species, mature fishes were collected. Different variety of fish species were collected for four different seasons. Bangladesh has a tropical monsoon climate with significant variations in rainfall and temperature throughout the country. There are four main seasons: (i) the pre-monsoon or summer during March-May, which has the highest temperatures;(ii) the monsoon or rainy-season during June-September, when the bulk of rainfall occurs; (iii) the post-monsoon or autumn during October-November ; (iv) the cool and sunny dry season or winter during December-February (Aquastat 2011).

In each season, samples were collected at the middle part of the season. During rainyseason, twenty-four fish species were collected on $30th$ July 2015; during autumn, twenty-two fish species were collected on $30th$ October, 2015; during winter, twentythree fish species were collected on $15th$ January, 2016 and during summer, twentytwo fish species were collected on $15th$ May, 2015. Twenty fish species were common in all four seasons.

Fish species were collected from the fishermen, just after caught by net on the bank of the river at 5 a.m.-7a.m. In case of each small fish, about 1kg fish were collected from five-six different fishermen and then mix together. In case of each large fish 3-5 fishes were collected from different fishermen. The fishes were kept in jip-locked plastic bag and then kept into chill-box at 4° C. The samples were transported to the laboratory immediately for further analysis. In the laboratory, at first the fish samples were identified by using the morphological characteristics, following Fishbase (2014), Rahman (2005) and Shafi and Quddus (1982). After identification, the measurements of some biological parameters (Length, width and weight) were taken and then the samples were kept in freezer at -20° C in jip-locked plastic bags with proper labeling of identification.

Fig. 2.4. Location of the sampling site in a map, Baidyer Bazar, Sonargaon Upazila

Table 2.1. Name of fish samples collected from Meghna River during rainyseason

Table 2.2. Name of fish samples collected from Meghna River during autumn

Table 2.3. Name of fish samples collected from Meghna River during winter

Table 2.4. Name of fish samples collected from Meghna River during summer

(a) Raining season (b) Autumn

Plate 2.1. Sampling site of the Meghna River in different seasons, a. Rainy-season, b. Autumn, c. Winter d. Summer season

Rui (*Labeo rohita*) Ghainna (*Labeo gonius*)

Bata (*Cirrhinus reba*) Jat punti (*Puntius sophore*)

Sharpunti (*Systomus sarana*) Tenga (*Mystus vittatus*)

Plate 2.2.a. Different fishes collected from the Meghna River

Bajaritengra (*Mystus tengra*) Gulsha (*Mystus cavaius*)

Shing (*Heteropnuestus fossilis*) Magur (*Clarias batrachus*)

Kachki (*Corica soborna*) Goldachingri (*M. rosenbergii*)

Plate 2.2.b. Different fishes collected from the Meghna River

Khalisha(*Trichogaster fasciata*) Chanda (*Parambassis ranga*)

Tarabaim (*Macrognathus aculiatus*) Borobaim (*Mastacembelus armatus*)

Chewa (*Pseudapocryptes elongatus*) Foli (*Notopterus notopterus*)

Plate 2.2.c. Different fishes collected from the Meghna River

Meni (*Nandus nadus*) Shole (*Channa striata*)

Gojar (*Channa marulius*) Bele (*Glossogobius. giuris*)

Poa (*Otolithoides. pama*) Bacha (*Eutropiichthys. vacha*)

2.10.3. Sample preparation

Before extraction the samples were taken out from the freezer and thaw well at room temperature. Then the scales, fins, viscera, gills were removed and washed with clean water. In case of small fish, whole body was grinded to paste and in case of large fishes, bones were removed and the remaining parts were grinded to paste with the help of the blender. The grinded fish samples were then ready for extraction.

2.11. Extraction of fish samples

The fish samples were extracted by Quick, Easy, Cheap, Effective, Rugged and Safe (QuEChERS) method on following Mastovska *et al.* 2010 with some modifications. 10 g of blended fish sample was taken in a Teflon centrifuge tube (50 mL) and ethylacetate (20 mL) was added and the content was vigorously shaken by hand for 1 min and vortexed for 1 min. Then anhydrous $MgSO_4$ (6 g) and NaCl (1.5 g) were added and the mixture was again vortexed for 1 min and then centrifuged (4000 rpm for 5 min). Supernatant (10 mL) was taken into a pre-weighted round bottom flask. Ethyle acetate was evaporated (below 40 °C) to dryness by rotary vacuum evaporator. Then the weight of the flask was taken. The lipid weight of the extracted fish sample was got by subtracting the initial weight of round bottom flask from final weight. After evaporation of solvent, re-dissolved in n-hexane (5 mL). The sample extract (2 mL of dissolved 5mL) was kept in a graduated test tube and then cleaned-up. The extraction procedure is demonstrated in Scheme-2.1.

2.12. Cleaned-up of fish samples

The 2mL extract in a graduated test tube was treated with 2 mL concentrated sulphuric acid (saturated with n-hexane). The test tube was vortexed for 1 minute and then centrifuge for 5 minutes. Supernatant (1 mL) was taken using pipette and kept into GC vial and analyzed by GC connected with an electron capture detector (ECD). The clean-up procedure is demonstrated in Scheme 2.1.

Scheme 2.1. Extraction and cleand-up method of Fish samples

1. Sample collection 2. Samples transported to Lab. immediately

3. Identification and take measurement 4. Samples kept in freezer

5. Sample preparation 6. Blended samples ready for extraction

Plate 2.3. Different steps prior to extraction

1. Extraction

2. Cleaned-up

3. Analysis by GC-ECD 4. Chromatogram of DDTs residues

Plate 2.4. Different steps of analysis of DDTs residues

2.13. Analysis of Organohalogen residues by GC-ECD

A Gas Chromatograph (GC-2010 Shimadzu) coupled with Electron Capture detector, (GC-ECD) was used for analysis. Separations were performed on HP-5 quartz capillary column (30 m long x 250 µm *i.d*.; 0.25 µm film thicknesses), nitrogen was used as carrier (column flow 1.92 mL/min.) as well as make up gas. The injector and detector temperatures were set at 220 °C and 290 °C, respectively. All injections were made in split-less/ split mode and injection volume was 1µL. The oven temperature was programmed as: initial temperature of 120 °C hold for 1 minute; increased at 20 °C min⁻¹ to 280 °C; hold for 4 min. Identifications of the organochlorine compounds analyte samples were done by comparing retention time of corresponding certified standard samples and quantification by using external calibration curves of the corresponding reference standard.

Table 2.5. Column oven temperature program

Initial Temperature : 120 °C

Fig. 2.5. Chromatogram of standards

Fig. 2.6. Chromatogram of control sample

Fig. 2.7.Chromatogram of a sample

2.14. Recovery, LOD and LOQ (Method Validation)

In order to make standard calibration curves for DDE, DDD, 2,4´-DDT and 4,4´-DDT stock solution of standard reference certified samples were serially diluted to obtain 12 different concentrations. The linearity of the method was well demonstrated over concentration range of 0.00025 -1.00 mg/L with an R^2 (regression coefficient) value in the range of 0.9840- 0.9960 (Table 2.6). The method was validated in terms of three replicates experiments. The percentage recoveries for fish samples were found to be 88-92 %, 101-113%, 76-104 % and 70-90 % for DDE, DDD, 2,4´-DDT and 4,4´- DDT respectively (Table 2.6), which were in the range 70-120 and acceptable for samples according to standard methodology. From the calibration curve the LOD and LOQ that are given in Table 2.6.

Residues	RT (min)	Linearity (R^2)	LOD	LOQ
DDE	10.363	0.9960	0.0625 ng g ⁻¹	0.2063 ng g ⁻¹
DDD	11.012	0.9950		
$2,4$ ^{-} DDT	11.099	0.9840		
$4,4$ ^{-} DDT	11.637	0.9960		

Table 2.6. Retention times (RT), Regression Coefficients (R^2) , Limit of Detection **(LOD), Limit of Quantification (LOQ) for DDTs in fish and prawn samples**

For recovery experiment control fish sample (Cultured Rui fish) was spiked separately with known amount of certified four standards at 3 different concentration levels (0.05, 0.10, 0.20 µg/mL or mg/kg). extraction and cleaned up were done following similar procedure as described in section 2.11 and 2.12. Present recovery and their RSD were calculated and presented in Table 2.7.

Fig. 2.8. Chromatogram of a control sample

Fig. 2.9. Chromatogram of standards in a recovery

2.15. Meteorological data of the sampling site

The accumulation of organohalogen residues in fishes has some relations to ambient temperature, rainfall and humidity. Therefore these meteorological data are needed to predict the influencing factors of DDTs accumulations in the present study. These data were collected from the Climate division, Meteorological department of Bangladesh. The monthly and seasonally data (March-December 2015, June-September 2015, October-November 2015 and January-February 2016) of Ambient bulk temperature, total rainfall and humidity of sampling site during sample collection were shown in Table 2.8.

Table 2.8. Average values of ambient temperature, rainfall, humidity of sampling site during the sample collection

2.16. Human health risk estimation

Health risk assessment of consumers from the intake of pesticides contaminated fish was characterized by using health risk index (HI). The estimated HIs were obtained by dividing the Estimated Daily Intake (EDI) by their corresponding values of Acceptable Daily Intakes (ADI) by WHO/FAO (FAO/WHO 2010) as shown by the equation;

 $HI = EDI / ADI$

When the HI is less than 1, the food concerned is considered acceptable. If it is greater than 1, the food concerned is considered a risk to the consumer (Darko and Akoto 2008, Akoto *et al.* 2015).

Estimated Daily Intake (EDI)

The EDIs of OCPs expressed as nanogram per kilogram body weight per day (ng/kg bw/d) were calculated as follows:

 $EDI = (C \times DR) / BW$

Where C is the measured concentration of OCPs ng g^{-1} , DR is average daily consumption rate of fish $(g day^{-1})$ and BW is body weight (kg).

2.17. Statistical analysis of data

Statistical analysis was performed using SPSS 22.0 software (SPSS Inc., Chicago, IL). One-way analysis of variance (ANOVA) followed by Tukey's HSD and LSD post hoc tests, were conducted for multiple comparison to test for significant differences between organohalogen residues of different seasons and different feeding modes. The relationship between residue levels with lipid contents, temperature, humidity and rainfall were assessed using pearson correlation. In all cases, the level of significance was set at 5% (*p*≤0.05).

Principle Component Analysis (PCA) was carried out on STATA 12 software to explore the relationship between residues and fishes of different modes. To identify the classification of species that indicates their similarity based on DDTs residues, Dendrogram for Agglomerative Hierarchial Clustering was performed using SPSS.

3. RESULTS AND OBSERVATIONS

3.1. Analysis of residual DDT and its metabolites in fish and prawn samples of different seasons from the Meghna River

Samples were collected periodically from four different seasons given in the Table 2.1- Table 2. 4. The samples were immediately wrapped with aluminium foil, put into a chill box with ice and transported to the laboratory on the same day. All the collected fish samples were identified and some morphometric data were taken. The samples were then stored in a freezer at a temperature below -20°C temperature until analysis. Before extraction, the fish tissue was made bone free, chopped and blended. The samples were extracted and cleaned up following the procedures described in Scheme 2.1, analyzed for the presence of residual DDT and its metabolites; DDE, DDD, 2,4**´** DDT and 4,4*´* DDT. The samples after cleaned up were analyzed by GC-ECD for the residual amounts of DDTs (DDT and its metabolites). The results of residual amounts of DDTs are given in the section 3.1.1 - 3.1.12.

3.1.1. Biological Parameters of fish and prawn samples for different seasons

Twenty-four, twenty-two, twenty-three and twenty-two different fishes and prawn species were studied during rainy-season, autumn, winter and summer respectively for the study. In each season only one species was prawn and remaing others were finfish species. In case of four different seasons, twenty species were found commonly to all seasons. In all seasons, the collected fish and prawn samples were mature and of maximum size. The length, width, weight, and sample number of samples of rainyseason, autumn, winter and summer are given in the Table 3.1, 3.2, 3.3 and 3.4 respectively.

In rainy-season, among the twenty-four analysed fish samples, the largest fish was the Boal (length was 70.49 cm, width was 12.88 cm) while the smallest fish was Kachki (length was 1.81 cm, width was 0.50 cm). The highest weight was recorded in Rui (1800.59 g) and lowest in Kachki (0.74 g). Three fishes were herbivore, nine fishes and one prawn were omnivore an eleven fishes were carnivore.

In autumn, among the twenty-two analysed fish samples, the largest fish was the Boal (length was 77.49 cm, width was 15.28 cm) while the smallest fish was Kachki (length was 1.92 cm, width was 0.72 cm). The highest weight was recorded in Rui (4050.69 g) and lowest in Kachki (0.74 g). Three fishes were herbivore, eight fishes and one prawn were omnivore and ten fishes were carnivore.

In winter, among the twenty-three analysed fish samples, the largest fish was the Boal (length was 72.49 cm, width was 15.28 cm) while the smallest fish was Kachki (length was 2.92 cm, width was 0.72 cm). The highest weight was recorded in Rui (1900.59 g) and lowest in Kachki (0.82 g). Three fishes were herbivore, eight fishes and one prawn were omnivore an eleven fishes were carnivore.

In summer, among the twenty-two analysed fish samples, the largest fish was the Boal (length was 76.89 cm, width was 12.98 cm) while the smallest fish was Kachki (length was 3.29 cm, width was 0.73 cm). The highest weight was recorded in Rui (4600.59 g) and lowest in Kachki (0.84 g). Three fishes were herbivore, seven fishes and one prawn were omnivore an eleven fishes were carnivore.

Table 3.1. Average length, width, weight and sample number of fish and prawn samples collected during rainy-season

No.	Common Name	Scientific Name	Length	Width	Weight	$n=$
			(cm)	(cm)	(g)	
$\mathbf{1}$	Rui	Labeo rohita	40.90	12.98	1800.59	$\mathbf{1}$
$\overline{2}$	Ghainna	Labeo gonius	30.48	8.38	870.00	$\overline{4}$
$\overline{3}$	Bata	Cirrhinus reba	14.63	2.73	30.11	10
$\overline{4}$	Jatpunti	Puntius sophore	7.57	3.37	10.49	$\overline{25}$
5	Sharpunti	Systomus sarana	22.33	6.53	144.24	$\overline{8}$
6	Tengra	Mystus vittatus	7.80	1.30	5.48	25
$\overline{7}$	Bajari-tengra	Mystus tengra	2.27	$\overline{1.00}$	1.33	40
$8\,$	Gulsha	Mystus cavaius	14.77	2.40	18.44	20
9	Shing	Heteropnuestes fossilis	17.13	2.70	37.40	20
$10\,$	Magur	Clarias batrachus	17.17	3.47	48.10	$\overline{15}$
11	Kachki	Corica soborna	1.80	0.50	0.74	250
12	Gutum	Lepidocephalus guntea	3.23	1.70	5.40	10
$\overline{13}$	Golda Chingri	Macrobrachium rosenbergii	26.98	4.27	179.68	35
14	$\overline{\text{Chanda}}$	Parmbassis ranga	4.43	1.67	2.71	20
15	Tara baim	Macrognathus aculiatus	4.83	1.83	17.27	25
16	Boro baim	Mastacembelus armatus	38.32	4.33	192.60	$\overline{5}$
17	Chewa	Pseudapocryptes elongatus	22.17	2.13	33.60	$\overline{15}$
18	Kaikka	Xenentodon cancila	19.73	1.53	61.31	$\overline{15}$
19	Foli	Notopterus notopterus	19.47	6.73	67.30	$\overline{8}$
20	Meni	Nandus nandus	14.27	4.47	50.71	15
21	Bele	Glossogobius giuris	24.30	5.23	147.99	$\overline{8}$
22	Poa	Otolithoides pama	23.20	5.27	121.47	10
23	Bacha	Eutropiichthys vacha	22.93	6.65	51.46	$\overline{6}$
24	Boal	Wallago attu	70.49	12.28	1002.22	$\overline{2}$

Table 3.2. Average length, width, weight and sample number of fish and prawn samples collected during autumn

Table 3.3. Average length, width, weight, and sample number of fish and prawn samples during winter

3.1.2. Lipid contents of fish and prawn samples of different seasons

3.1.2.1. Lipid contents during rainy-season

In rainy-season, the lipid contents (%) of different fish and prawn species ranged from $0.53 \pm 0.01\%$ in Boal to $13.98 \pm 1.50\%$ in Bacha (Table 3.5). Considering the amount of lipid the chronology is Boal < Kaikka <Golda-chingri < Kachki < Bele < Poa < Meni < Shing < Chewa < Foli < Chanda < Borobaim < Tarabaim < Bajari-tengra < Ghainna < Gutumn < Gulsha < Bata < Sharpunti < Tengra < Jatpunti < Magur < Rui. Among the analyzed fishes, Gulsha, Bata, Sharpunti, Tengra, Jatpunti, Magur, Rui and Bacha contained higher amount of lipid $(4.67 \pm 0.29\%$ to $13.98 \pm 1.50\%$) while Chanda, Barabaim, Tarabaim, Bajari-tengra, Ghainna and Gutumn contained medium amount of lipid (2.65 \pm 0.05 % to 4.56 \pm 0.11 %) and Boal, Kaikka, Golda-chingri, Kachki, Bele, Poa, Meni, Shing, Chewa and Foli contained lower amount of lipid (0.53 ± 0.03) % to 1.98 ± 0.07 %) are shown in Fig. 3.1 where red coloured bar reported for higher, green for medium and blue for lower amount of lipid.

Fig. 3.1. Bar diagram of lipid content (%) of fish and prawn samples from the Meghna River during rainy-season

Table 3.5. Lipid content (%) of fish and prawn samples from the Meghna River during rainy-season

No.	Name	Scientific name	Lipid $(\%)$				
			$Mean \pm SD$	RSD			
$\mathbf{1}$	Rui	Labeo rohita	13.27 ± 0.64	4.85			
$\overline{2}$	Ghainna	Labeo gonius	4.52 ± 0.41	9.09			
3	Bata	\overline{Cirrhi} nus reba	5.68 ± 0.27	3.07			
4	Jat punti	Puntius sophore	7.94 ± 0.22	2.8			
5	Sharpunti	Systomus sarana	5.71 ± 0.25	4.34			
6	Tengra	Mystus vittatus	7.03 ± 0.26	3.76			
$\overline{7}$	Bajari-tengra	Mystus tengra	4.08 ± 0.23	12.00			
8	Gulsha	Mystus cavasius	4.67 ± 0.29	6.28			
9	Shing	Heteropnuestes fossilis	1.91 ± 0.01	0.21			
10	Magur	Clarias batrachus	9.17 ± 1.47	16.02			
11	Kachki	Corica soborna	1.01 ± 0.15	15.2			
12	Gutum	Lepidocephalus guntea	4.56 ± 0.11	2.32			
13	Golda Chingri	Macrobrachium rosenbergii	0.90 ± 0.07	7.63			
14	Chanda	Parambassis ranga	2.65 ± 0.05	1.71			
15	Tara baim	Macrognathus aculiatus	3.53 ± 0.13	3.62			
16	Boro baim	Mastacembelus armatus	2.71 ± 0.17	6.14			
17	Chewa	Pseudapocryptes elongatus	1.95 ± 0.03	0.42			
18	Kaikka	Xenentodon cancila	0.53 ± 0.10	2.67			
19	Foli	Notopterus notopterus	1.98 ± 0.07	3.34			
20	Meni	Nandus nandus	1.73 ± 0.15	8.73			
21	Bele	Glossogobius giuris	1.13 ± 0.02	1.69			
22	Poa	Otolithoides pama	1.65 ± 0.01	0.43			
23	Bacha	Eutropiichthys vacha	13.98 ± 1.50	9.63			
24	Boal	Wallago attu	0.53 ± 0.03	5.75			

*SD = Standard Deviation, RSD= Relative Standard Deviation

3.1.2.2. Lipid contents during autumn

In autumn season, the mean lipid contents (%) of different fish and prawn samples ranged from $0.51 \pm 0.06\%$ in Boal to $14.43 \pm 0.64\%$ in Rui (Table 3.6) and the increased in the following order : Goldga-chingri < Shole < kachki < Chanda < Bele < Chewa < Foli < Poa < Shing < Meni <Bajari-tengra < Ghainna < Khalisha < Borobaim < Bata < Gulsha < Tengra < Jatpunti < Sharpunti < Bacha < Rui. Among the analyzed fishes, Borobaim, Bata, Gulsha, Tengra, Jatpunti, Sharpunti, Bacha, Rui contained higher amount of lipid $(5.52 \pm 0.16\%$ to $14.43 \pm 0.64\%$) while Foli, Poa, Shing, Meni, Bajari-tengra, Ghainna, Khalish contained medium amount of lipid (2.30 ± 0.30 % to 4.95 ± 0.70 %) and Boal, Goldachingri, Shole, Kachki, Chanda, Bele and Chewa contained lower amount of lipid $(0.51 \pm 0.06\%$ to 2.01 $\pm 0.09\%$) are shown in Fig. 3.2 where red coloured bar reported for higher, green for medium and blue for lower amount of lipid.

Fig. 3.2. Bar diagram of lipid content (%) of fish and prawn samples from the Meghna River during autumn

*SD = Standard Deviation, RSD= Relative Standard Deviation

3.1.2.3. Lipid contents during winter

In the winter, the mean lipid contents (%) of different fish and prawn species ranged from 0.41 ± 0.07 in Shole to $13.66 \pm 0.56\%$ in Rui (Table 3.7.). Considering the value the chronology is Shole < Chewa < Boal < Golda-chingri < Tarabaim < Chanda < Kachki < Foli < Bele < Shing < Poa < Borobaim < Meni < Bajari tengra < Ghainna < Gulsha < Tengra < Magur < Bata < Sharpunti < Jatpunti < Bacha < Rui. In the analyzed fishes Rui, Bacha, Jatpunti, Sharpunti, Bata, Magur, Tengra, Gulsha contained higher amount of lipid $(6.07 \pm 0.28\%$ to $13.66 \pm 0.56\%$), Ghainna, Bajaritengra, Meni and Borobaim contained medium amount (3.49 \pm 0.21% to 4.60 \pm 0.39%) and Poa, Shing, Bele, Foli, Kachki, Chanda, Tarabaim, Goldachingri, Boal, Chewa and Shole contained lower amount of lipid $(0.41 \pm 0.07\%$ to $2.39 \pm 0.56\%)$ are shown in Fig. 3.3 where red coloured bar reported for higher, green for medium and blue for lower amount of lipid.

Fig. 3.3. Bar diagram of lipid content (%) of fish and prawn samples from the Meghna River during winter

Table 3.7. Lipid contents (%) of fish and prawn samples from the Meghna River during winter

*SD = Standard Deviation, RSD= Relative Standard Deviation

3.1.2.4. Lipids contents during summer

In summer season, the mean lipid contents (%) of different fish and prawn species ranged from $0.12 \pm 0.01\%$ in Boal to $15.57 \pm 0.01\%$ in Bacha (Table 3.8.). Considering the value the chronology is Boal < Gojar < Shole < Golda chingri < Foli < Chewa < Bele < Kachki < Chanda < Shing < Borobaim < Poa < Meni < Ghainna < Gulsha < Sharpunti < Tengra< Bajari-tengra < Bata < Jatpunti < Rui < Bacha. Among the analyzed fishes, Gulsha, Sharpunti, Tengra, Bajari-tengra, Jatpunti, Rui, Bacha contained higher amount of lipid $(6.67 \pm 0.13 \%$ to $15.57 \pm 0.14 \%)$ while Chanda, Shing, Borobain, Meni, Ghainna contained medium amount of lipid $(2.88 \pm 0.01 \%)$ to $4.59 \pm 0.03\%$) and Boal, Gojar, Shole, Goldachingri, Foli, Chewa, Bele and Kachki contained lower amount of lipid $(0.12 \pm 0.01\% -1.99 \pm 0.15\%)$ shown in Fig. 3.4 where red coloured bar reported for higher, green for medium and blue for lower amount of lipid.

Fig. 3.4. Bar diagram of lipid content (%) of fish and prawn samples from the Meghna River during summer

Table 3.8. Lipid content (%) of fish and prawn samples from the Meghna River during summer

*SD = Standard Deviation, RSD= Relative Standard Deviation

3.1.3. DDT and its metabolite residues in four different seasons

3.1.3.1. DDT and its metabolites during rainy-season

The mean concentrations of DDE, DDD, 2,4**´** DDT and 4,4**´** DDT residues in the fishes and prawn species during rainy-season are presented in the Table 3.9.

The concentrations of DDE residue ranged from 0.58 ± 0.01 ng g⁻¹ in Ghainna to 10.32 \pm 0.83 ng g⁻¹ in Meni. The concentrations of DDD residue ranged from 0.71 \pm 0.09 ng g^{-1} in Bajari-tengra to 170.83 \pm ng g^{-1} .in Bacha. The mean concentrations of 2,4**´**-DDT residue ranged from BDL (Below detection limit) in Poa to 21.65 ± 2.81ng g^{-1} in Gulsha. The concentrations of 4,4²-DDT residue ranged from 0.17 \pm 0.05 ng g^{-1} in Kaikka to 22.11 \pm 0.87 ng g⁻¹ in Meni. The compositional profile of DDTs in fish and prawn samples showed the contribution of 4,4**´**-DDT (3.44-58.49%), 2,4**´**-DDT (4.85-89.37%), DDD (0.001-72.04%) and DDE (2.17-64.25%) during rainy-season. DDD (55.54%) was the major contributor in fish and prawn samples in this season and followed by $4,4$ -DDT (16.10%), DDE (14.36%) and $2,4$ -DDT(14.00%) (Fig. $3.5.$).

3.1.3.2. DDT and its metabolites during autumn

The mean concentrations of DDE, DDD, 2,4**´** DDT and 4,4**´** DDT residues in twentytwo different fish and prawn species during autumn are presented in the Table 3.10.

The concentrations of DDE residue ranged from 4.22 ± 0.39 ng g⁻¹ in khalisha to 164.14 \pm 4.21 ng g⁻¹ in Bacha. The concentrations of DDD residue ranged from 3.27 \pm 0.63 ng g⁻¹ in Shing to 53.87 \pm 5.39 ng g⁻¹ in Bacha. The value of 2,4²-DDT residue ranged from BDL in Bata to 31.19 \pm 3.13 ng g⁻¹in Bacha. The concentrations of 4,4²-DDT residue ranged from 0.06 \pm 0.003 ng g⁻¹ in Bata to 22.32 \pm 0.86 ng g⁻¹in Bacha. The compositional profile of DDTs in fish and prawn samples showed the contribution of 4,4**´**-DDT(15.12-89.29%), 2,4**´**-DDT(8.15-56.97%), DDD (0.00- 34.97%) and DDE (4.18-22.70%) In autumn season, the major contributor is DDE (53.11%) and followed by the DDD (31.79%) , $2,4$ -DDT (8.52%) and $4,4$ -DDT(6.58%). (Fig. 3.6.).

Fig. 3.5. Composition profile (%) of DDTs in fish and prawn samples during rainy-season

Fig. 3.6. Composition profile (%) of DDTs in fish and prawn samples during autumn

Number	Name	Scientific name	DDE		DDD		$2.4'$ -DDT		$4.4'$ - DDT		Σ DDTs	
			$Mean \pm SD$	RSD	$Mean \pm SD$	RSD	$Mean \pm SD$	RSD	$Mean \pm SD$	RSD	$Mean \pm SD$	RSD
	Rui	Labeo rohita	0.62 ± 0.02	3.75	2.45 ± 0.20	8.25	3.34 ± 0.21	5.88	7.08 ± 0.02	0.24	13.48 ± 0.07	0.50
$\overline{2}$	Ghainna	Labeo gonius	0.58 ± 0.01	1.78	1.47 ± 0.13	9.01	1.89 ± 0.17	9.41	7.08 ± 0.20	0.24	11.02 ± 0.13	1.14
3	Bata	Cirrhinus reba	2.27 ± 0.08	3.62	1.89 ± 0.27	14.32	1.46 ± 0.14	9.78	1.03 ± 0.05	4.39	6.65 ± 0.40	6.06
$\overline{4}$	Jat punti	Puntius sophore	2.62 ± 0.17	6.35	1.26 ± 0.06	4.48	4.42 ± 0.15	3.32	4.09 ± 0.48	11.80	12.39 ± 0.63	5.06
5	Sharpunti	Systomus sarana	4.86 ± 0.09	1.87	4.95 ± 0.41	8.17	1.11 ± 0.15	13.41	0.58 ± 0.12	20.31	11.50 ± 0.33	2.84
6	Tengra	Mystus vittatus	2.26 ± 0.12	5.22	0.85 ± 0.12	13.75	2.51 ± 0.41	16.15	1.93 ± 0.26	13.51	7.55 ± 0.87	11.53
τ	Bajari-tengra	Mystus tengra	2.38 ± 0.07	2.91	0.71 ± 0.09	13.43	2.29 ± 0.06	2.62	1.84 ± 0.14	7.87	7.22 ± 0.36	5.03
8	Gulsha	Mystus cavasius	8.37 ± 2.42	28.9	23.69 ± 6.21	26.20	21.65 ± 2.81	12.97	18.62 ± 4.16	2.12	72.33 ± 7.47	10.32
9	Shing	Heteropnuestes fossilis	10.08 ± 0.87	8.62	45.86 ± 2.83	6.18	5.87 ± 0.94	16.08	9.12 ± 1.45	15.91	70.93 ± 2.43	3.43
10	Magur	Clarias batrachus	5.65 ± 0.92	16.22	1.91 ± 0.35	18.29	2.51 ± 0.25	10.13	2.72 ± 0.34	12.39	12.79 ± 1.21	9.49
11	Kachki	Corica soborna	1.55 ± 0.08	4.99	0.31 ± 0.21	0.31	0.41 ± 0.03	6.73	0.38 ± 0.05	12.79	2.64 ± 0.35	13.38
12	Gutum	Lepidocephalus guntea	3.11 ± 0.24	7.76	6.94 ± 0.10	1.42	0.52 ± 0.12	6.04	0.31 ± 0.08	13.80	10.87 ± 0.51	3.39
13	Golda Chingri	Macrobrachium rosenbergii	1.18 ± 0.11	9.06	1.06 ± 0.02	2.25	1.59 ± 0.33	20.66	1.72 ± 0.15	8.54	5.55 ± 0.29	3.87
14	Chanda	Parambassis ranga	3.07 ± 0.07	2.39	4.36 ± 0.11	2.55	0.57 ± 0.01	0.48	0.76 ± 0.01	0.96	8.76 ± 0.03	0.34
15	Tara baim	Macrognathus aculiatus	5.25 ± 1.11	11.53	12.9 ± 2.02	15.62	3.52 ± 0.35	9.92	0.84 ± 0.16	19.19	22.52 ± 2.63	10.20
16	Boro baim	Mastacembelus armatus	7.55 ± 0.09	11.91	3.57 ± 0.19	3.58	9.76 ± 2.03	20.80	11.66 ± 0.90	7.70	32.54 ± 3.72	11.42
17	Chewa	Pseudapocryptes elongatus	6.74 ± 0.65	9.66	36.36 ± 3.29	9.07	1.83 ± 0.10	5.69	3.36 ± 0.54	15.97	48.29 ± 4.12	8.59
18	Kaikka	Xenentodon cancila	1.72 ± 0.12	7.07	1.73 ± 0.48	4.73	0.38 ± 0.60	16.12	0.17 ± 0.05	8.89	4.00 ± 0.61	4.59
19	Foli	Notopterus notopterus	1.75 ± 0.42	16	1.83 ± 0.12	6.44	1.19 ± 0.05	3.91	0.69 ± 0.01	1.62	5.45 ± 0.34	3.01
20	Meni	Nandus nandus	10.32 ± 0.83	8.07	58.3 ± 2.01	3.44	8.53 ± 1.19	13.94	22.11 ± 0.87	3.93	99.23 ± 2.75	2.77
21	Bele	Glossogobius giuris	7.72 ± 1.05	13.67	18.46 ± 3.11	16.84	9.13 ± 1.66	19.76	14.45 ± 1.51	10.48	49.77 ± 4.04	8.12
22	Poa	Otolithoides pama	8.02 ± 0.06	0.79	9.09 ± 0.40	4.38	nd		0.38 ± 0.01	3.18	17.50 ± 0.65	3.66
23	Bacha	Eutropiichthys vacha	6.58 ± 1.02	15.44	170.83 ± 29.31	17.16	7.05 ± 0.67	4.99	6.68 ± 0.43	6.51	191.14 ± 31.18	16.31
24	Boal	Wallago attu	2.17 ± 0.04	1.92	0.82 ± 0.15	19.13	12.19 ± 0.21	9.46	1.74 ± 0.19	11.13	16.91 ± 0.54	7.77
Range			$0.58 - 10.32$		$0.71 - 170.83$		nd - 21.65		$0.17 - 22.11$		$2.64 - 191.14$	

Table 3.9. Level of different organohalogen pesticide residues of fishes collected during rainy season (March-August, 2015) from the Meghna River. (Values express as mean \pm SD in ng g⁻¹ in f.w., n = 3 replicates)

*f.w.- Fresh weight, nd- Non detectable, SD-Standard deviation, RSD- Relative standard deviation

No.	Name	Scientific name	DDE		DDD		$2.4'$ - DDT		$4.4'$ - DDT		Σ DDTs	
			$Mean \pm SD$	RSD	$Mean \pm SD$	RSD	$Mean \pm SD$	RSD	$Mean \pm SD$	RSD	$Mean \pm SD$	RSD
$\overline{1}$	Rui	Labeo rohita	41.28 ± 6.04	14.62	3.77 ± 0.74	19.5	0.74 ± 0.10	13.67	0.44 ± 0.06	13.25	46.24 ± 6.75	14.60
$\overline{2}$	Ghainna	Labeo gonius	25.85 ± 0.99	3.84	4.60 ± 0.38	8.16	0.70 ± 0.08	11.06	0.62 ± 0.11	17.55	31.78 ± 0.83	2.60
3	Bata	Cirrhinus reba	27.64 ± 2.62	9.49	4.00 ± 0.24	6.1	nd		0.06 ± 0.003	5.54	31.70 ± 2.87	9.01
$\overline{4}$	Jat punti	Puntius sophore	57.06 ± 0.77	1.34	10.94 ± 0.25	2.33	0.65 ± 0.05	7.36	0.86 ± 0.17	19.50	69.50 ± 0.62	0.89
5	Sharpunti	Systomus sarana	50.79 ± 2.73	5.38	8.81 ± 1.12	12.7	0.48 ± 0.10	20.58	0.46 ± 0.10	20.99	60.54 ± 4.04	6.67
6	Tengra	Mystus vittatus	64.80 ± 17.08	26.35	13.23 ± 1.47	11.1	0.66 ± 0.13	20.36	0.60 ± 0.12	15.90	79.28 ± 18.38	23.18
τ	Bajari-tengra	Mystus tengra	43.91 ± 2.27	5.16	6.04 ± 0.38	0.38	0.49 ± 0.03	5.18	0.13 ± 0.01	11.38	50.57 ± 2.67	5.28
8	Gulsha	Mystus cavasius	53.56 ± 8.46	15.80	6.89 ± 0.68	9.85	0.64 ± 0.08	12.88	0.40 ± 0.09	21.62	61.48 ± 8.78	14.28
9	Shing	Heteropnuestes fossilis	19.36 ± 2.52	13.03	3.27 ± 0.63	19.18	0.30 ± 0.06	18.76	0.17 ± 0.02	11.61	23.11 ± 1.98	8.56
10	Kachki	Corica soborna	37.51 ± 2.24	5.96	4.98 ± 0.25	5.11	0.51 ± 0.08	14.83	0.17 ± 0.02	12.18	43.16 ± 2.01	4.67
11	Golda chingri	Macrobrachium rosenbergii	44.37 ± 4.16	9.38	6.20 ± 0.95	15.3	0.79 ± 0.11	14.26	0.77 ± 0.01	1.19	52.13 ± 3.48	6.68
12	Khalisha	Trichogaster fasciata	4.22 ± 0.39	9.28	8.81 ± 1.16	13.2	1.31 ± 0.03	2.17	2.08 ± 0.34	16.55	16.42 ± 1.90	11.57
13	Chanda	Paarambassis ranga	34.15 ± 3.05	8.93	6.09 ± 0.96	15.8	0.39 ± 0.04	10.03	0.40 ± 0.02	6.10	41.03 ± 4.03	9.81
14	Boro baim	Mastacembelus armatus	49.88 ± 4.75	9.52	6.26 ± 0.67	10.6	0.22 ± 0.03	13.60	0.12 ± 0.12	12.01	56.49 ± 5.41	9.58
15	Chewa	Pseudapocryptes elongatus	46.06 ± 3.98	8.64	7.04 ± 0.64	9.14	0.73 ± 0.01	13.27	0.40 ± 0.04	9.42	54.24 ± 4.72	8.70
16	Foli	Notopterus notopterus	51.19 ± 6.02	11.80	5.48 ± 1.60	20	0.78 ± 0.11	14.32	0.82 ± 0.01	1.19	58.26 ± 7.29	12.50
17	Meni	Nandus nandus	63.30 ± 2.29	4.62	10.46 ± 1.33	12.71	0.83 ± 0.15	17.85	0.50 ± 0.12	24.50	75.08 ± 4.45	5.93
18	Shol	Channa striata	8.79 ± 2.19	24.91	20.13 ± 4.73	20	18.38 ± 3.94	21.41	10.83 ± 2.26	20.81	58.13 ± 8.14	14.01
19	Bele	Glossogobius giuris	63.85 ± 6.99	10.80	14.31 ± 1.35	9.46	0.66 ± 0.09	15.37	0.44 ± 0.01	1.57	80.12 ± 7.82	9.76
20	Poa	Otolithoides pama	53.47 ± 3.62	6.78	7.39 ± 0.92	12.09	0.78 ± 0.09	12.09	0.42 ± 0.04	9.18	62.06 ± 4.18	6.73
21	Bacha	Eutropiichthys vacha	164.14 ± 4.21	2.56	53.87 ± 5.93	11.01	31.19 ± 3.13	10.03	22.32 ± 0.86	3.84	271.5 ± 6.17	2.27
22	Boal	Wallago attu	11.33 ± 0.74	6.54	18.95 ± 0.69	3.67	25.02 ± 4.53	18.10	16.24 ± 2.09	12.86	71.55 ± 7.70	10.76
Range		$4.22 - 164.14$		$3.27 - 53.87$		nd-31.19		0.06-22.32		$16.42 - 271.5$		

Table 3.10. Level of different organohalogen pesticide residues of fishes collected during autumn (October-November 2015) from the Meghna River. (Values express as mean ± SD in ng g-1 in f.w., n=3 replicates)

*f.w.- Fresh weight, nd- Non detectable, SD-Standard deviation, RSD- Relative standard deviation

3.1.3.3. DDT and its metabolites during winter

The mean concentrations of DDE, DDD, 2,4^{\prime} DDT and 4,4^{\prime} DDT residues in twentythree different fish and prawn species during winter are presented in the Table 3.11.

The concentrations of DDE ranged from 1.83 \pm 0.06 ng g⁻¹ in Tarabaim to 48.16 \pm 5.99 ng g^{-1} in Bele. The concentrations DDD ranged from BDL in Goldachingri to 81.42 ± 15.01 ng g⁻¹ in Bacha. The mean concentrations 2, 4² DDT ranged from 0.16 \pm 0.07 ng g⁻¹ng/g in Boal to 19.65 \pm 3.94 ng g⁻¹ in Borobaim. The concentrations 4, 4² DDT ranged from 0.13 ± 0.01 ng g⁻¹ in Shole to 13.76 ± 2.28 ng g⁻¹ in Bacha.

The compositional profile of DDTs in fish and prawn samples showed the contribution of 4,4**´**-DDT(12.22-97.05%), 2,4**´**-DDT (0.001-65.15%), DDD (0.87- 38.75%) (Fig. 3.7.) In winter the major contributor is DDE (68.80%) and followed by DDD(18.23%), 2,4 - DDT(7.63%) and 4,4 - DDT(5.33%).

3.1.3.4. DDT and its metabolites summer

The mean concentrations of DDE, DDD, 2,4**´** DDT and 4,4**´** DDT residues in twentytwo different fish and prawn species during summer are presented in the Table 3.12.

The concentrations of DDE ranged from 20.54 \pm 3.34 ng g⁻¹ in Boal to 476.53 \pm 48.16ng g^{-1} in Bacha. The concentrations of DDD ranged from 17.22 \pm 0.45 ng g^{-1} in Shing to 329.71 \pm 4.08 ng g⁻¹ in Bacha. The concentrations of 2,4^{\degree}DDT ranged from 45.0 ± 5.80 ng g⁻¹ in Kachki to 427.22 ± 3.63 ng g⁻¹ in Poa. The concentrations of 4,4 DDT ranged from 55.49 \pm 2.28 ng g⁻¹ in Shing to 625.39 \pm 32.88 ng g⁻¹ in Gojar.

The compositional profile of DDTs in fish and prawn samples showed the contribution of 4,4**´**-DDT(4.73-29.05%), 2,4**´**-DDT(8.25-31.21%), DDD (11.93- 40.85%) and DDE (29.51-65.12%) (Fig. 3.8). In this season, the major contributor is 4,4**´**-DDT(40.23%) and followed by 2,4**´**-DDT(27.59%), DDE (18.60%) and DDD (13.58%).

Fig. 3.7. Composition profile (%) of DDTs in fish and prawn samples during winter

Fig. 3.8. Composition profile (%) of DDTs in fish and prawn samples during summer

No.	Name	Scientific name	DDE		DDD		$2.4'$ -DDT		$4,4'$ -DDT		Σ DDTs	
			$Mean \pm SD$	RSD	$Mean \pm SD$	RSD	$Mean \pm SD$	RSD	$Mean \pm SD$	RSD	$Mean \pm SD$	RSD
$\overline{1}$	Rui	Labeo rohita	4.61 ± 0.36	4.89	5.81 ± 0.28	4.88	4.01 ± 0.77	18.80	3.92 ± 0.75	19.06	18.35 ± 2.06	11.20
2	Ghainna	Labeo gonius	3.37 ± 0.43	12.89	4.97 ± 0.36	7.25	3.92 ± 0.35	9.001	3.16 ± 0.32	10.23	15.43 ± 0.78	5.04
$\overline{3}$	Bata	Cirrhinus reba	5.51 ± 0.49	15.51	10.46 ± 0.32	3.04	0.21 ± 0.07	7.16	0.30 ± 0.05	17.19	16.48 ± 0.82	4.73
$\overline{4}$	Jat punti	Puntius sophore	5.99 ± 0.32	5.28	10.10 ± 1.09	10.78	0.31 ± 0.04	11.15	0.22 ± 0.05	21.02	16.62 ± 1.35	8.11
5	Shar punti	Sytomus sarana	5.06 ± 0.16	3.15	9.40 ± 0.26	2.78	0.29 ± 0.01	1.33	0.19 ± 0.02	9.24	14.94 ± 0.36	2.39
6	Tengra	Mystus vittatus	6.13 ± 0.22	6.13	12.19 ± 0.32	2.63	0.21 ± 0.07	6.78	0.30 ± 0.01	1.77	18.83 ± 0.44	2.41
7	Bajari-tengra	Mystus tengra	8.08 ± 1.48	18.34	12.00 ± 1.78	14.84	1.22 ± 0.07	5.88	1.69 ± 0.42	24.65	22.98 ± 3.19	13.88
8	Gulsha	Mystus cavasius	7.06 ± 0.89	12.58	$\overline{10.99} \pm 1.29$	11.8	0.99 ± 0.29	18.78	0.25 ± 0.02	9.50	19.28 ± 1.96	10.53
9	Shing	Heteropnuestes fossilis	2.11 ± 0.38	18.19	0.90 ± 0.02	2.76	0.32 ± 0.05	16.14	0.88 ± 0.16	17.85	4.21 ± 0.32	7.68
10	Magur	Clarias batracus	6.29 ± 0.19	2.99	2.13 ± 0.22	10.36	0.85 ± 0.11	12.49	0.70 ± 0.09	12.52	9.97 ± 0.24	2.31
-11	Kachki	Corica soborna	3.87 ± 0.27	6.85	6.85 ± 0.48	7.03	0.17 ± 0.01	6.69	0.24 ± 0.05	8.27	11.13 ± 0.82	7.34
12	Golda Chingri	Macrobrachium rosenbergii	3.23 ± 0.66	18.65	nd		0.34 ± 0.04	10.68	0.31 ± 0.03	8.81	3.88 ± 0.60	15.52
13	Chanda	Paraambassis ranga	6.03 ± 0.61	10.05	5.36 ± 0.36	6.67	0.75 ± 0.04	5.55	1.42 ± 0.03	1.82	13.57 ± 0.85	6.30
14	Tara Baim	Macrognathus aculiatus	1.83 ± 0.06	3.01	1.22 ± 0.11	8.69	5.93 ± 0.31	5.21	6.00 ± 0.51	8.56	14.98 ± 0.88	5.90
15	Boro baim	Mastacembelus armatus	8.35 ± 0.99	11.86	13.10 ± 1.98	5.14	19.65 ± 3.94	20.05	9.61 ± 1.31	4.64	50.70 ± 5.24	10.33
16	Chewa	Pseudapocryptes elongatus	4.16 ± 0.50	11.98	12.17 ± 2.89	23.75	6.04 ± 1.01	16.70	2.52 ± 0.40	13.61	24.90 ± 2.01	8.07
17	Foli	Notopterus notopterus	3.23 ± 0.43	13.31	6.75 ± 0.32	4.67	0.17 ± 0.01	2.70	0.21 ± 0.04	16.87	10.36 ± 0.68	6.61
18	Meni	Nandus nandus	9.01 ± 1.47	16.36	4.90 ± 0.09	1.77	0.88 ± 0.06	6.28	0.71 ± 0.06	7.89	15.49 ± 1.37	8.83
19	Shol	Channa striata	26.30 ± 4.22	16.06	0.25 ± 0.06	23.42	0.42 ± 0.04	10.57	0.13 ± 0.01	4.64	27.11 ± 4.31	15.88
20	Bele	Glossogobius giuris	48.16 ± 5.99	12.44	5.57 ± 1.02	18.29	1.17 ± 0.11	9.60	0.43 ± 0.06	14.21	55.33 ± 6.97	12.59
21	Poa	Otolithoides pama	39.67 ± 4.68	11.79	3.77 ± 0.68	18.01	0.52 ± 0.07	13.16	0.37 ± 0.05	12.58	44.33 ± 4.77	10.77
22	Bacha	Eutropiichthys vacha	34.06 ± 3.39	9.96	81.42 ± 15.01	18.38	12.33 ± 1.77	14.32	13.76 ± 2.28	16.59	141.57 ± 10.24	7.23
23	Boal	Wallago attu	16.68 ± 0.06	3.60	1.45 ± 0.21	14.31	0.16 ± 0.003	1.94	0.14 ± 0.02	13.01	18.44 ± 0.76	4.10
Range		$1.83 - 48.16$		nd-81.42		$0.16 - 19.65$		$0.13 - 13.76$		3.88 141.57		

Table 3.11. Level of different organohalogen pesticide residues of fishes collected during winter (December 2015-February 2016) from the Meghna River. (Values express as mean ± SD in ng g-1 in f.w., n= 3 replicates)

*f.w.- Fresh weight, nd- Non detectable, SD-Standard deviation, RSD- Relative standard deviation

No.	Name	Scientific name	DDE		DDD		$2.4'$ -DDT		$4.4'$ -DDT		Σ DDTs	
			$Mean \pm SD$	RSD	$Mean \pm SD$	RSD	$Mean \pm SD$	RSD	$Mean \pm SD$	RSD	$Mean \pm SD$	RSD
	Rui	Labeo rohita	32.01 ± 3.71	11.58	41.42 ± 0.65	1.56	102.28 ± 2.95	2.88	74.68 ± 8.97	12.01	250.40 ± 9.17	3.66
2	Ghainna	Labeo gonius	25.81 ± 2.28	3.84	41.42 ± 5.11	11.06	97.81 ± 2.80	2.87	69.10 ± 2.14	3.10	234.15 ± 0.65	0.28
3	Bata	Cirrhinus reba	43.90 ± 2.69	6.14	57.02 ± 2.51	4.41	66.96 ± 4.86	7.26	141.53 ± 10.72	7.58	309.42 ± 20.36	6.58
$\overline{4}$	Jat punti	Puntius sophore	68.86 ± 8.81	12.79	36.72 ± 5.22	14.23	94.35 ± 6.69	7.09	85.78 ± 15.83	18.45	285.71 ± 15.30	5.36
5	Sharpunti	Systomus sarana	49.49 ± 2.33	4.70	38.45 ± 6.03	15.68	93.23 ± 5.10	5.47	88.02 ± 12.08	13.72	269.19 ± 15.81	5.87
6	Tengra	Mystus vittatus	138.01 ± 4.72	3.42	49.86 ± 1.61	3.23	133.81 ± 4.59	3.43	153.44 ± 3.94	2.57	475.11 ± 11.67	2.46
$\overline{7}$	Bajari-tengra	Mystus tengra	92.59 ± 3.24	3.50	66.42 ± 4.29	6.47	163.54 ± 19.67	12.03	259.90 ± 5.401	2.08	582.44 ± 32.29	5.54
8	Gulsha	Mystus cavasius	31.42 ± 6.42	20.43	68.55 ± 12.07	17.60	161.32 ± 18.82	11.67	143.83 ± 11.83	8.23	405.12 ± 48.42	11.95
9	Sing	Heteropnuestes fossilis	40.86 ± 3.87	9.48	17.22 ± 0.45	2.59	55.88 ± 1.78	3.19	55.49 ± 2.28	4.12	169.45 ± 8.11	4.79
10	Kachki	Corica soborna	25.91 ± 0.52	2.02	23.23 ± 1.45	6.24	45.01 ± 5.80	12.88	63.42 ± 2.77	4.37	157.58 ± 1.15	0.73
11	Golda-chingri	Macrobrachium rosenbergii	30.65 ± 0.24	0.77	54.93 ± 0.24	0.44	74.79 ± 1.07	1.44	172.70 ± 1.21	0.70	333.07 ± 2.70	0.81
12	Chanda	Parambassis ranga	25.91 ± 4.88	18.84	21.84 ± 1.70	7.80	47.25 ± 4.73	10.01	63.42 ± 2.77	4.37	158.43 ± 5.72	3.61
13	Boro Baim	Mastacembelus armatus	185.84 ± 8.19	4.40	142.60 ± 18.90	13.25	309.20 ± 7.28	2.35	480.60 ± 7.19	1.50	1118.24 ± 34.74	3.11
14	Chawa	Pseudapocryptes elongatus	46.82 ± 1.57	3.35	116.07 ± 2.24	1.93	84.73 ± 2.48	2.93	462.36 ± 8.41	1.82	709.98 ± 11.07	1.56
15	Foli	Notopterus notopterus	10403 ± 2.25	2.16	37.11 ± 1.56	4.2	111.35 ± 6.103	5.48	$\overline{168.23} \pm 4.70$	2.79	420.72 ± 13.96	3.32
16	Meni	Nandus nandus	173.01 ± 5.93	3.43	94.87 ± 8.42	8.88	233.95 ± 18.27	7.80	372.91 ± 45.03	12.07	874.74 ± 31.15	3.56
17	Shol	Channa striata	210.31 ± 8.25	3.92	63.76 ± 3.14	4.92	219.46 ± 4.93	2.25	279.07 ± 14.87	5.33	772.59 ± 26.21	3.39
18	Gojar	Channa marulius	226.20 ± 11.24	4.97	150.92 ± 1.19	0.79	419.98 ± 36.69	8.74	625.39 ± 32.88	5.26	1422.48 ± 77.83	5.47
19	Bele	Glossogobius giuris	34.61 ± 1.55	4.47	121.09 ± 1.13	0.94	50.45 ± 4.24	8.41	181.84 ± 1.80	0.99	387.99 ± 8.52	2.20
20	Poa	Otolithoides pama	292.78 ± 1.95	0.67	113.16 ± 5.70	5.04	427.22 ± 3.63	0.85	510.60 ± 3.60	0.71	1343.77 ± 6.43	0.48
21	Bacha	Eutropiichthys vacha	476.53 ± 48.16	10.11	329.71 ± 4.08	1.24	357.86 ± 70.30	19.6	496.80 ± 62.16	12.51	1660.89 ± 157.9	9.51
22	Boal	Wallago attu	20.54 ± 3.34	16.28	48.23 ± 6.73	13.96	174.74 ± 17.14	9.81	190.74 ± 8.81	4.62	443.24 ± 35.72	8.23
Range		$20.54 - 476.53$		$17.22 - 329.71$			$45.01 - 427.22$		$55.49 - 625.39$		$157.58 - 1660.89$	

Table 3.12. Level of different organohalogen pesticide residues of fishes collected during summer (March - May 2015) from the Meghna River. (Values express as mean ± SD in ng g-1 in f.w., n = 3 replicates)

*f.w.- Fresh weight, nd- Non detectable, n-3, Replicate number is three, SD-Standard deviation, RSD- Relative standard deviation

3.1.4. Total DDTs residues of fishes and prawn of different seasons

3.1.4.1. Total DDTs (∑DDTs) residues during rainy-season

The mean concentrations of total DDTs (Σ DDTs) residue (ng g⁻¹ ww) in fishes and prawn during rainy-season are shown in Table. 3.13. In this season **∑**DDTs ranged from 2.64 \pm 0.35 ng g⁻¹in Kachki to 191.14 \pm 31.18 ng g⁻¹ in Bacha and increased in the following order : Kachki < Kaikka < Foli < G.Chingri < Bata < Boal < Bajaritengra < Tengra < Chanda < Gutumn < Ghainna < Sharpunti < Jatpunti < Magur < Rui < Tarabaim < Poa < Borobaim < Chewa < Bele < Shing < Gulsha< Meni < Bacha. For fish consumption, European Union (Binelli and Provini, 2004) has established a Maximum admissible concentration (MAC) of 50 ng g^{-1} ww for **∑**DDTs. Considering the valueof MAC, Gulsha, Shing, Meni, Bele and Bacha contained higher amount of Σ DDTs (>50 ng g⁻¹ ww) while the remaining others contained lower (<50 ng g^{-1} ww) amount are shown in Fig. 3.9 where brown bar represents higher and green for lower amount.

Fig. 3.9. Bar diagram of total DDTs (∑DDTs) residues in fish and prawn samples from the Meghna River during rainy-season

3.1.4.2. Total DDTs (∑DDTs) residues in autumn

The mean concentrations of total DDTs (Σ DDTs) residue (ng g⁻¹ ww) in fishes and prawn during autumn are shown in Table. 3.14. **∑DDTs** ranged from 16.42 ± 1.90 ng g^{-1} in Kalisha to 271.50 \pm 6.17 ng g^{-1} in Bacha and increased in the following order : Kalisha < Shing < Bata < Ghainna < Chanda < Kachki < Rui < Bajaritengra < Goldachingri < Chewa < Borobaim < Shole < Foli < Sharpunti < Gulsh < Poa < Jatpunti < Boal < Meni < Tengra < Bele < Bacha. The **∑**DDTs residues in most of the samples during autumn were higher than rainy-season. Jatpunti, Sharpunti, Tengra, Bajari-Tengra, , Gulsha, Goldachingri, Borobaim,Chewa, Foli, Meni, Shole, Bele, Poa, Boal and Bacha contained higher amount of Σ DDTs residues (>50 ng g⁻¹) while Rui, Ghainna, Bata, Shing, Kachki, Khalisha and Chanda contained lower (<50 ng g- $¹$) shown in Fig.3.10 where brown bar represents higher and green for lower amount.</sup>

Fig. 3.10. Bar diagram of total DDTs (∑DDTs) residues in fish and prawn samples from the Meghna River during autumn

Table 3.14. Total DDTs (∑DDTs) residues in fish and prawn samples from the Meghna River during autumn (values express as mean ± SD in ng g -1 in w. w.)

3.1.4.3. Total DDTs (∑DDTs) residues during winter

The mean concentrations of total DDTs (Σ DDTs) residue (ng g⁻¹ ww) in fishes and prawn during winter are shown in Table. 3.15. ∑DDTs ranged from 3.88 ± 0.60 ng g⁻¹ in Golda Chingri to 141.57 ± 10.24 ng g⁻¹ in Bacha and considering the residue levels the chronology is Gloda chingri \langle Sing \langle Magur \langle Foli \langle Kachki \langle Chanda \langle Sharpunti < Tarabaim < Ghainnna < Meni < Jatpunti < Bata < Rui < Boal < Tengra < Gulsha < Bajari-tengra <Chewa < Shole < Poa < Bacha. The ∑DDTs residues in the overall samples in winter lower than rainy-season and autumn. Among the analyzed samples, Borobaim, Bele and Bacha contained higher amount of **∑**DDTs (>50 ng g-1 ww) while the remaining others contained lower amount $(<50$ ng g^{-1} ww) are reported in Fig 3.11, brown coloured bar represents higher and green for lower amount.

Fig. 3.11. Bar diagram of total DDTs (∑DDTs) residues in fish and prawn samples from the Meghna River during winter season

Table 3.15. Total DDTs (∑DDTs) residues in fish and prawn samples from the Meghna River during winter (values express as mean ± SD in ng g-1 in w. w.)

3.1.4.4. Total DDTs (∑DDTs) residues during summer

The mean concentrations of total DDTs (Σ DDTs) residue (ng g⁻¹ ww) in fishes and prawn during summer are reported in Table. 3.16. Σ DDTs ranged from 157.58 \pm 1.15 ng g⁻¹ in Kachki to 1660.89 \pm 157.9 ng g⁻¹ in Bacha and increased in the following : Kachki < Chanda < Shing < Ghainna < Sharpunti < Rui < Jatpunti < Bata < Goldachingri < Bele < Gulsha < Foli < Boal < Tengra < Bajari-tengra < Chewa < Shole < Meni < Borobaim < Poa < Gojar < Bacha. The DDTs residues in all the samples during summer were much higher (> 50 ng g⁻¹) than other three seasons are shown in Fig.3.12.

Fig. 3.12. Bar diagram of total DDTs (∑DDTs) residue in fish and prawn samples from the Meghna River during summer

Table 3.16. Total DDTs (∑DDTs) residues in fish and prawn samples from the Meghna River during summer (values express as mean ± SD in ng g -1 in w. w.)

3.1.5. Total mean of lipid contents and total DDTs concentrations of four seasons Among different analyzed fish and prawn species, twenty speceies were common for all seasons. In case of that twenty species, the average lipid contents of four seasons are shown in Table 3.17 and Fig.3.13.a. Rui exhibited highest lipid content with an average of 14.19% and Boal exhibited lowest content with an average of 0.44 %. Considering these values the chronology is Boal $(0.44) <$ Goldachingri $(1.03) <$ Bele (1.33) < Chewa (1.42) < Kachki (1.68) < Foli (1.82) < Chanda (2.18) < Poa (2.65) < Shing (2.69) < Meni (3.55) < Borobaim (3.79) < Bujuri-Tengra (5.39) < Gulsha (5.87) < Sharpunti (6.08) < Tengra (6.99) < Bata (7.03) < Jatpunti (9.18) < Bacha (13.93) < Rui.(14.19).

The average total ∑DDTs concentrations of four different seasons of the commonly found twenty fishes are shown in Table 3.17 and Fig. 3.13.b. Bacha exhibited the highest ∑DDTs concentration with an average of 634.16 ng g^{-1} while Kachki exhibited the lowest with an average of 53.63 ng g^{-1} . Considering these values the chronology is Kachki $(53.63 \text{ ng } g^{-1})$ < Chanda (55.45) < Shing (66.92) < Ghainna (73.11) < Rui (82.17) < Sharpunti (89.11) < Bata (91.10) < Jatpunti (96.11) < Goldachingri (98.66) < Foli (123.71) < Boal (135.04) < Gulsha (139.55) < Bele (143.31) < Tengra (145.20) < Bajari-tengra (165.81) < Chewa (209.35) < Meni (266.14) < Borobaim (314.50) < Poa (366.92) < Bacha $(634.16$ ng g⁻¹).

Fig. 3.13. Bar diagram of (a) mean of total lipid contents and (b) mean of total ∑DDTs residue in fishes and prawn sampls during four different seasons

Table 3.17. Average lipid content and ∑DDTs residues of fish and prawn samples of different seasons

Fig. 3.15. Hierarchical dendrogram for fishes and prawn of summer season represented by total DDTs residues

The hierarchical cluster analysis (Dendogram)

The hierarchical cluster analysis (Ward's method applying Pearson correction) of fishes and prawn species using average linkage between groups and square Euclidean distance which uncover the similarities or dissimilarities between species depending on organohalogen residues. Two cluster analyses was done, one for the fishes commonly found in all seasons and second for the fishes of summer season as contained maximum amount of DDTs residues (Fig. 3.14. and Fig. 3.15 respectively).

The dendogram of Fig. 3.14. grouped the species into three big clusters with subgroups. The first cluster contains all herbivores and the omnivores that mostly feed on plant; Kachki, Chanda, Ghainna, Shing, Bata, Sharpunti, Jatpunti, Gurachingri and Rui. The second cluster contains omnivores that mostly feed on animals and the carnivores of low fat; Bajari-tengra, Chewa, Tengra, Bele, Gulsha, Boal and Foli. The third cluster contains bottom feeder and highly carnivorous; Borobaim, Meni, Poa and Bacha.

The dendogram of Fig. 3.15. grouped the species into four clusters with subgroup. The first cluster contains herbivores and omnivores of mostly plant feeder; Kachki, Chanda, Shing, Bata, Gurachingri, Rui, Ghainna, Jatpunti and Sharpunti. The second cluster contains omnivores of mostly animal feeder and carnivores with low fat; Foli, Boal, Gulsha, Bele, Tengra and Bajari-terngra. The third cluster contains highly carnivores; Chewa, Shole and Meni. The fourth cluster contains also highly carnivores and bottom feeder Gojar, Poa, Borobaim and Bacha.

3.1.7. Total DDTs (∑DDTs) concentrations in herbivore, omnivore and carnivore

During rainy-season, the concentrations of total DDTs residue in herbivore fishes ranged from of 11.02 ± 0.13 ng g⁻¹ in Ghainna to 13.48 ± 0.07 ng g⁻¹ in Rui, in case of omnivores the ranged from 2.64 \pm 0.35 ng g⁻¹ in Kachki to 72.33 \pm 7.47 ng g⁻¹ in Gulsha and in case of carnivores the range was 4.00 ± 0.61 ng g⁻¹ in Kaikka to 191.14 \pm 31.81 ng g⁻¹ in Bacha (Fig. 3.16.a.). The mean concentration of total DDTs for all the herbivores, omnivores and carnivores in this season were 10.38, 21.38 and 45.10 ng g^{-1} .respectively.

During autumn, in case of herbivore fishes the concentrations of total DDTs residue ranged from 31.70 \pm 2.87 ng/g in Bata to 46.24 \pm 6.75 ng g⁻¹ in Rui, in case of omnivores the range was 16.42 ± 1.90 ng g⁻¹ in kalisha to 79.28 \pm 18.38 ng g⁻¹ in Tengra and in case of carnivores the range was 41.03 ± 4.03 ng g⁻¹ in Meni to 271.50 \pm 6.71 ng g⁻¹ in Bacha (Fig. 3.16.b.). The mean concentration of total DDTs for all the herbivores, omnivores and carnivores in this season were 36.37, 47.34, 96.00 ng g⁻ ¹.respectively.

During winter, in case of herbivore fishes the concentrations of total DDTs residue ranged from 15.43 \pm 0.78 ng g⁻¹ in Ghainna to 18.35 \pm 2.06 ng g⁻¹ in Rui, in case of omnivores the range was 3.88 \pm 0.60 ng g⁻¹ in Goldachingri to 22.98 \pm 3.19 ng g⁻¹ in Bajari-Tengra and in case of carnivores the range was 10.36 ± 0.68 ng g⁻¹ in Fole to 141.57 ± 10.24 ng g⁻¹ in Bacha (Fig.3.17.a.). The mean concentration of total DDTs for all the herbivores, omnivores and carnivores in this season were 16.75, 20.32 and 43.30 ng g^{-1} . respectively.

During summer, in case of herbivore fishes the concentrations of total DDTs residue ranged from 234.15 \pm ng g⁻¹ in Bata to 309.42 \pm ng g⁻¹ in Ghainna, in case of omnivores the range was $157.58 \pm$ ng g⁻¹ in Kachki to 582.44 \pm ng g⁻¹g in Tengra and in case of carnivores the range was $158.43 \pm$ ng g⁻¹ Chanda to $1660.89 \pm$ ng g⁻¹ in Bacha (Fig.3.17.b.). The mean lipid contents for all the herbivores, omnivores and carnivores in this season were 264.64, 334.71, 845.83 ng/g.respectively.

One way ANOVA with Tukey HSD and LSD tests showed significant differences in total DDTs between herbivore, omnivore and carnivore fishes $(f = 3.098, p < .050)$.

Fig. 3.16. Bar diagram of total DDTs residue in fish and prawn samples from the Meghna River during (a) rainy- season and (b) autumn

Fig. 3.17. Bar diagram of total DDTs residue in fish and prawn samples from the Meghna River during (a) winter and (b) summer

3.1.8. Multivariate statistical analysis (Principle Component Analysis-PCA)

On the basis of correlation matrix, a Euclidean biplot was obtained using the first two axis of PCA (Fig. 3.18.). The first axis explained 68.10% of total variance and second axis accounted for 24.6% of variance.

Fig. 3. 18. PCA on the basis of feeding habit of fish species.

3.1.9. Correlation between DDTs residues with lipid, rainfall, temperature and humidity

Correlation analysis was performed in order to check for the existence of correlations between DDTs residues, lipid and some meteorological parameters (temperature, humidity and rainfall). In all cases, the level of significance was set at 5% ($p \le 0.05$). The results of Pearson's correlation coefficients are illustrated in Table 3.18, Apendix 17. and Fig. 3.19 - 3.21. The analysis revealed significant ($p \le 0.01$) positive correlations between DDE and DDD with lipid contents (e.g. lipid vs DDE- $r = .210^*$, p<0.05; lipid vs DDD - $r = 275**$, p<0.01). Highly significant relationship shown between temperature and humidity with DDT and its derivatives. Temperature showed significant positive relationship with DDE, DDD, 2,4**´**-DDT, 4,4**´**- DDT and ∑DDTs residues (r=0.243*, p=0.02; r=0.338**, p=0.001; r=0.379**, p= 0.000; r=0.380 ** , p=0.000 and r=0.368 ** , p=0.000 respectively). Wherese there were highly significant negative relationship between humidity and DDTs, where $r = -0.318**$, p=0.002; r= $-0.276**$, p= 0.008; r= $-0.273**$, p=0.009 and r= $-0.285**$, p=0.006 for the correlation of DDE, 2,2**´**-DDT, 4,4**´**-DDT and total DDTs (∑DDTs) with humidity respectively. DDTs residues correlated with rainfall as negatively but not significant.

$\frac{1}{2}$. The state of the state of $\frac{1}{2}$									
	Rainfall.	Temperature	Humidity	Lipid	DDE	DDD	2,4'-DDT	4,4'-DDT	Σ DDTs
Rainfall		$.684**$.955**	$-.047$	-187	.010	$-.060$	$-.057$	$-.080$
Temperature		1 $\bf{1}$	$.525**$.040	$.243*$	$.338*$	$.379**$	$.380**$	$.368**$
Humidity			1	$-.070$	$-.318**$	$-.161$	$-.276**$	$-.273**$	$-.285**$
Lipid				1	$.210*$	$.275**$.090	038	.129
DDE						$.774**$	$.835***$	$.785***$	$.897**$
DDD						1	$.756**$	$.791**$	$.861**$
$2,4'$ -DDT								$.940**$	$.967**$
$4,4'$ -DDT									$.970**$

Table 3.18. Pearson correlation matrix between DDTs, lipid and some meteorological parameters

** Correlation is highly significant at the level 0.01 level * correlation is significant at the level 0.05 level.

∑DDTs 1

Fig. 3.19. Correlation between ∑DDTs residues in fishes with temperature

Fig. 3.20. Correlation between ∑DDTs residues in fishes with humidity

Fig. 3.21. Correlation between ∑DDTs residues in fishes with rainfall

3.1.10. Seasonal variations of lipid contents and DDTs concentrations

Total DDTs (∑DDTs) concentraitons vary significantly between species for all seasons $(p<0.01)$. One way ANOVA with Tukey HSD and LSD tests showed highly significant differences in total DDTs concentrations of fishes and prawn between seasons $(p<0.01)$ where the summer season showed highly significant differences between rainy-season, autumn and winter $(p<0.01$ in all cases; summer vs rainyseason-p=0.000, summer vs autumn-p=0.000 and summer vs winter-p=0.000) while there were no significant difference in ∑DDTs concentrations among rainy-season, autumn and winter $(p>0.05$ in all cases; rainy-season vs autumn-p=1.00, rainy-season vs winter-p=0.555, autumn vs winter-p=.559). The lipid contents in fishes and prawn showed variation between seasons but there is no significant differences (p>0.05).

In case of lipid content, 55% species showed the highest lipid contents in summer and lowest in rainy-season. The mean values of total lipid contents of all samples for each seasons shown in Fig. 3.22.a. Considering these values the chronology is rainy-season (4.28%) < winter (4.14%) < autumn (4.95%) < summer $(5.46\%).$

In case of ∑DDTs concentrations 70% species showed the highest amount in summer and lowest in rainy-season. The mean values of total ∑DDTs concentrations in all samples for the seasons shown in Fig. 3.22.b. Considering these values the chronology is winter $(25.60 \text{ ng g}^{-1})$ < rainy-season $(30.88 \text{ ng g}^{-1})$ < autumn $(63.34 \text{ ng g}^{-1})$ g^{-1}) < summer (580.72 ng g^{-1}).

(a)

(b)

Fig. 2.23.a. Chromatogram of a sample in rainy season

Fig. 2.23.b. Chromatogram of a sample in autumn

Fig. 2.23.c. Chromatogram of a sample in winter

Fig. 2.23.d. Chromatogram of a sample in summer

3.1.11. Recent or historical use of DDT

The (DDE+DDD)/DDTs ratio can be used to establish whether its input occurred recently or in the past, and also whether degradation of DDT is significant or not. If the ratios were less than 0.5, DDT can be used as fresh input instead of degraded as historical resides (Yu *et al.*2011). In this study, there were 38.46% of the total samples of four seasons having values of (DDE+DDD)/DDTs lower than 0.5 and 61.54% having values upper than 0.5. (Fig. 3.24.). The values of the (DDE+DDD)/DDTs for fishes of four seasons are given in the Table 3.19.

Fig. 3.24. Ratios of (DDE+DDD)/DDTs) in different fish and prawn samples from the Meghna River

Table 3.19. Ratios of (DDE+DDD)/DDTs) in different fish and prawn samples

3.1.12. Human health risk

Health risk assessment of consumers from the intake of pesticides contaminated fish was characterized by using Health risk Index (HI). The estimated HIs were obtained by dividing the Estimated Daily Intake (EDI) by their corresponding values of Acceptable Daily Intakes (ADI) by WHO/ FAO (2010) as shown by the equation;

$HI = EDI / ADI$

When the HI is less than 1, the food concerned is considered acceptable. If it is greater than 1, the food concerned is considered a risk to the consumer (Darko and Akoto 2008; Akoto *et al.* 2015).

Estimated Daily Intake (EDI)

The EDIs of OCPs expressed as nanogram per kilogram body weight per day (ng/kg bw/d) were calculated as follows:

 $EDI = (C \times DR) / BW$

Where C is the measured concentration of OCPs ng g^{-1} , DR is average daily consumption rate of fish (g day⁻¹) and BW is body weight (kg) which was set at 60 kg (WHO, 2010).The average daily consumption rate was derived form fisheries resource survey of Bangladesh, 2013-14 and the value is 52.88 g/day (NFW 2015). The values of EDI of fishes and prawn samples of four seasons are given in the Table 3.20.

Acceptable Daily Intake (ADI)

ADI represents the daily concentration below which there is a high probability of no adverse health effect. It is an estimate of the residue that can be ingested by a person daily over an extended period of time without suffering deleterious effects. ADI is expressed by body mass per kilogram per day which was set at 10000 ng /kg bw/d

The hazard indices presented in Table 3.21. All the detected residues of DDTs were <1 .

Table 3.20. EDI of different fish and prawn samples

Table 3.21. HI of different fish and prawn samples

4. DISCUSSION

4.1. DDT and metabolites in fish and prawn samples

The chemical analyses showed that DDT and its derivatives were detected in variable quantities in all species through out the year indicating their widespread contamination in the aquatic environment of Meghna river system. The potential route and possible sources of DDT contamination of aquatic ecosystems are the surface runoff as a result off the possible illegal use of the pesticides in surrounding areas and also atmospheric deposition (Bouwman *et al.* 2008, Bornman *et al.* 2007). Therefore, the wide detection of DDTs in analysed fish and prawn species may be related to the extensive applications of DDT in the surrounding environments. According to the district statistics huge amount of food grains, vegetables, banana trees were cultivated on the bank of the Meghna River at Sonargaon Upazila (NDS 2013). So there may be possible illegal use of DDT as pesticide in crop production on the bank of the river which could be washing out into the adjacent waterbody and DDT residues were evident from the analysed fishes. Reports from fishes of other river and common fish markets also indicate higher level of DDT residues (Nahar *et al.* 2008, Hossain *et al.* 2016).

Significant differences of DDTs levels were found among species in each season. The results imply that the bioconcentration of DDTs in fish is species-specific due to their ecological characteristics such as feeding habits and habitats. Moreover lipid content, dietary consumption, metabolism rate and excretion rate are all the factors of primary importance in explaining body burden of DDTs. Considering the average concentrations of total DDTs (∑DDTs) residue of four seasons, the twenty fish and prawn species that analysed in all seasons showed the chronology of Kachki (*Corica soborna*) < Chanda (*Parambassis ranga*) < Shing (*Heteropnuestes fossilis)* < Ghainna *(Labeo gonius*) < Rui (*Labeo rohita*) < *Systomus sarana* (Sharpunti) < Bata (*Cirrhinus reba*) < Jatpunti (*Puntius sophore*) < Goldachingri (*Macrobrachium rosenbergii*) < Foli (*Notopterus notopterus*) < Boal (*Wallago attu*) < Gulsha (*Mystus cavasius*) < Bele (*Glossogobius giuris*) < Tengra (*Mystus vittatus*) < Bajari-tengra

(*Mystus tengra*) < Chewa (*Pseudapocrypter elongates*) < Meni (*Nandus nandus*) < Borobaim (*Mastacembelus armatus*) < Poa (*Otolithoides pama*) < Bacha (*Eutropiichthys vacha*). The other fishes that analysed in one or two seasons showed the chronology, Kaikka (*Xenentodon cancila*) < Khalisha (*Trychogaster fasciata*) <*Gutum (Lepidocephalus guntea*) < Magur (*Clarius batrachus*) < Tarabaim (*Macrognatus aculiatus*) < Shole (*Channa striata*) < Gojar (*Channa marulius*). Therefore, among these twenty-seven analyzed species, Rui (*L. rohita*), Ghainna (*L. gonius*), Bata (*C. reba*), Jatpunti (*P. sophore*), Sharpunti (*S. sarana*), Shing (*H. fossilis*), Magur (*C. batrachus*), Kachki (*C. soborna*), Goldachingri (*M. rosenbergii*), Gutum (*L. guntea*), Khalisha (*T. fasciata*), Kaikka (*X. cancila*), Tarabaim (*M. aculeatus*) and Boal (*W. attu*) conained lower amount of DDTs residue. While Tengra (*M. vittatus*), Bajari-tengra (*M. tengra*), Gulsha (*M. cavasius*), Borobaim (*M. armatus*), Chewa (*P. elongatus*), Meni (*N. nandus*), Bele (*G. giuris*), Poa (*O. pama*) and Bacha (*E. vacha*) tissue contained higher amont of total DDTs residues.

Trophic position and lipid content of aquatic organisms are reliable predictors of OCP concentrations in aquatic ecosystem (Kidd *et al.* 2000, Crosly *et al.* 1998). The present analysis showed that Kachki, Chanda and Shing fish contained low amount of lipid contents. Kachki fish mainly feed on phytoplankton and zooplankton while Shing feed on insects and plant materials due to their omnivorous nature (Shafi and Quddus 1982). Chanda is a carnivorous fish feed on zooplankton and larvae. As these three fishes occupy lower position in the food chain (may be just after the herbivore) and together with their low lipid contets may be related to their low DDTs residues.

Rui, Ghainna and Bata fish mainly feed on phytoplankton and algae due to their herbivorous nature (Fishbase 2014, Shafi and Quddus 1982). The present study showed that these fishes contained higher amount of lipid. Although containing higher lipid, they occupy the second trophic level just after the producer may be responsible for containing lower amount of DDTs residue. Similarly herbivore fish *Sarotherofon galilaeus* contained lower concentrations of organochlorine residues as it feed on phytoplankton and algae (Akoto *et al.* 2016).

Jatpunti and Sharpunti mainly feed on algae, plant material and small amount insects due to omnivorous nature (Fishbase 2014, Shafi and Quddus 1982) while contained higher lipid contents may be responsible for comperatively lower DDTs accumulation but higher than Kachki, Shing, Magur, Chanda, Rui, Ghainna and Bata.

Only prawn species is Goldachingri contained lower amount of DDTs residue may be related to their low lipid content. Moreover Goldachingri is an omnivorous species mainly feed on plankton, Diatoms, Copepods and small crustaceans. Similar result was reported by Nahar *et al.* (2008).

Foli (*N. notopterus*) and Boal (*W. attu*) also contained lower amount of DDTs residues. Foli is a carnivore fish feed on insects, prawns, nematods, aquatic weed and bottom sands (Kiran and Waghray, 1996). Although it occupy in the higher trophic level but contained low amount of lipid content may be responsible for lower amount of DDTs residues. Boal is a highly carnivorous fish but contained low DDTs residues as it contained very low amount of lipid content moreover it's exretion rate is high. According to ASTRD 2002 DDT leaves the body mostly in Urine. So there may be relation between it's exscretion rate, lipid content and low DDTs residue. Similar findings repoted in the Boal fish from othe rivers and common fish market (Hossain *et al.* 2016, Nahar *et al.* 2008). Deribe *et al.* 2011 also reported that African big barb (*Barbus itermedius)* is positioned at the higher trophic level but contained low concentration of DDTs compare to others at the same trophic level due to its lower relative lipid content.

It is well established that the extent of accumulation of organochlorine compounds is greater in fish that have high lipid content (Muir *et al.* 1990). Because of lipophilic nature, organohalogen compounds can biomagnify and bioaccumulate in carnivore fish which as they are generally positioned at the top of trophic level in an ecosystem (Deribe *et al.* 2011, Zhou *et al.* 2007). Tengra, Bajari-tengra and Gulsha are omnivorous fishes feed mainly on small fishes, insects, moluscks and little amount of algae and plant material (Chaklader *et al.* 2014, Gupta *et al.* 2014) together with higher amount of lipid contents may be the cause of containing higher DDTs residues.

Chewa (*P. elongatus*) depends on shrimps and non-shrimp crustaceans mainly include copepods, crab larvae, mysids and amphipods predominantly as it is carnivore (Rahman *et al.* 2016). Meni (*N. nadus*) is a bottom and column feeder and feed mainly on small fish, prawn, fish fry, chironomid and insect larvae and predominantly carnivorous fish (Mustafa *et al.* 1980). Bele (*G. giuris*) is carnivore and cannibalistic in nature. Food items mainly consist of fish, crustaceans, insects, zooplankton and on the other hand considerable time of the year a recognizable proportion of food composed of juvenile of bele (Hossain *et al.* 2016). Poa (*O. pama*) is highly carnivore fish with predatoy in nature in both juvenile and adult life. It feeds mainly on small fishes, fish fry and bottom dwelling invertebrates (Manojkumar and Acharia 1990). The Shole (*C. striata*) and Gojar (*C. marulius*) feed on small sized fish, frog, snake, insects, earthworm and tadpole larva due to their highly carnivorous and predatory in nature (Amin *et al.* 2014).

From the above discussion on the food and feeding habits of these carnivorous fish species, it is clear that these fishes depend on diverse food items cover several trophic levels which make them as top consumers in the aquatic body. According to ATSDR (2002) and Connell (1995), persistent lipophilic organic compounds bioaccumulate and biomagnify with increasing trophic levels. Similar trends of DDT accumulation to higher levels in fishes of higher trophic levels observed, reaching levels thousand times higher than in water and organisms at lower trophic level. Therefore, the higher amount of DDTs residues in Chewa, Meni, Shole, Gojar, Bele, Poa, Shole and Gojar in the present study are in accordance with their higher trophic position in food chain. Similarly the significant higher concentration of DDTs were also found in carnivorous species African catfish (*Clarias gariepinus)* from the lake Koka, Ethoipia, marine fish snakefish from Natuna Island, South China sea and fresh water catfish (*Clarias anguillaris)* from Tono reservoir of Ghana (Derib *et al.* 2011, Hao *et al.* 2014 and Akoto *et al.* 2016). Similarly the carnivore and predator fishes like silver catfish, (*Schilbe intermedius)* in Tono reservoir of Ghana, contained higher concentration of organochlorine residues (Akoto *et al.* 2016).

Borobaim (*M. armatus*) contained higher amount of DDTs which may be attributed as it is a bottom-dwelling carnivorous habits and inhabits in muddy tunnel which may increase its exposure to DDTs residue at the bottom sediment. After entering into the aquatic habitat, DDT settle down and deposited into the bottom sidiments may be related to the higher DDT accumulation in sediment feeder fishes. Similar finding, relatively higher concentrations of organochlorine pesticide residues were also found in bottom-feeding species in Oueme River catchment in the Repablic of Benin (Pazou *et al*. 2006) and bottom-feeding marine fish from China and the United States (Xia *et al*. 2012, Morgan and Lohmann 2010, Sun *et al*. 2014). Kent and Johnson (1979) also reported that bottom feeder catfish, *Clarius anguillaris* contained highest level of OC pesticides in the Americal Fall Reservoir. However, data from the National Contaminant Biomonitoring Program in major US Rivers and Great lakes found no differences between OC pesticide residues in bottom feeders and predatory fish. (Caldas *et al*. 1999).

From the present analysed fish and prawn species, Bacha fish (*E. vacha*) contained highest amount of DDTs residues in all seasons. Several factors may be responsible for its higher DDTs accumulation. Bacha contained much higher amount of lipid content. According to Shafi and Quddus (1982), Abbas (2010), Gupta and Benergi (2016), in adult stage, Bacha fish is a carnivorous and piscivorous fish feed on small fishes, juvenile of taki and other large fishes, mollusks, crustaceans and annelids etc. Their feeding intensity is also higher as they are voracious. Therefore the high lipid content together with piscivorous and voracious nature of Bacha may be responsible for its higher DDTs acuumulation in tissue. Connell (1995) observed the same findings and concluded that fatty carnivorous fish species are expected to have higher concentration of POPs than lean fish in the same trophic level.

4.2. Composition profiles of DDT and its metabolites

When deposition in water, DDT will bind to particles in water, settle and be deposited in the sediment. Then, it is taken up by small organisms and fish (ASTRD 2002). After deposition in water, 4, 4ʹ- DDT and 2,4ʹ- DDT are broken down to DDD and DDE by sunlight and by the micro-organisms (Ssebugere *et al*. 2009, Rao *et al.* 2014). From this study, the maximum percentage of 4,4́ DDT (40.2% of the sum of DDT) reported in summer season followed by rainy-season, winter and autumn, this may indicting the recent input of DDT prior to summer season. While DDD and DDE showed lower percentage in summer and rainy-season but increasing to the next two seasons. It could be happened that during summer, DDT deposited to the water-body as before i.e. during winter season huge amount of vegetables cultivated on the bank of the river where pesticide may be used that persist in the soil which could be runoff in the river through the short summer rain season. After introduce into the waterbody the DDTs then enter the food chain, enters into the fish body and converted to DDD and DDE within next three seasons. Micro-organisms and oxygenase enzymes present in the fish body may be responsible from rapid convertion of 4,4ʹ- DDT to DDD and DDE. As we know, metabolic functions through oxygenase (enzyme) in different organisms specially in fish has a central role in converting 4,4ʹ- DDT into DDE (Muralidharan *et al*. 2009). According to Binelli and Provini 2003 and Kongwick *et al*. 2006, DDT metabolis faster in fish where its half life is approximately eight months in fish.

4.3. Understanding differences in DDTs residues among fishes of different feeding habits and lipid contents

A wide range of DDTs concentrations was observed in different fishes. As stated earlier in countries other than Bangladesh, the lipid content and trophic positions influence the accumulation of contaminatant concentrations in fishes (Singh and singh 2008, Arnot and Gobas 2004, Canbana and Rusmussen 1994, Pastor *et al.*1996). This study focused on the DDTs concentration in relation to the trophic position of fishes and also relationship to their lipid contents. In this respect, statistical analysis showed significant differences in total DDTs between fishes of different trophic levels; herbivore, omnivore and carnivore fishes. Moreover Pearson correlation between lipid contents with DDT and its derivatives showed significant direct positive correlations between lipid contents with DDE and DDD other than DDT. It could be attributed that although DDT is lipophilic but in the fish body DDT rapidly breakdown to DDE and DDD and then stored to fish fat. Similar relationships between DDE and lipid content was found in *Schilbe marmoratus* fish in Congo river basin, reported by Verhaet *et al*. (2013).

Through out the seasons, the mean DDTs residual amount for different trophic positioned fishes; herbivores, omnivores and carnivores showed the chronology as carnivore > omnivore > herbivore. The DDTs accumulation found to be increased with higher trophic levels. Similar bioaccumulation of organochrine compounds reported in plankivore fish Enjunguri (*Haprochromis nigripinnis*) from lake Edward , Uganda displayed no detectable levels of DDTs (Ssebugere *et al.* 2009), in omnivore Silver fish (*Rastreneobola argentea*) from lake Victoria and lake Kyoga in Uganda displayed lower DDTs residues (Nnamuyomba *et al.,* 2014) and in carnivore Tiger fish (*Hydrocynus vittatus)* from the Luvuvhu River of South Africa displayed the highest OCP bioaccumulation (Gerber *et al.* 2014). Similarly, Verhaert *et al.* (2013) also reported significant increase of DDT with increasing trophic levels in cases fishes from Itaimbiri, Aruwimi and Lomami river basins.

Some exception is that all species of higher trophic levels did not contained higher residue levels. The omnivore fishes, Kachki and Shing contained residual levels lower than that of some herbivore species while the carnivore fishes, Chanda and Boal contained residue levels lowere than that of some omnivore species. This may be related to their lower lipid contents that mentioned before. Similar findings was reported carnivore fish Boal fish of common fish market contained low amount of DDTs residues (Nahar *et al.* 2008). On the otherhand the herbivore fish Rui contained much higher lipid content but its residue level was lower which could be attributed to its lower trophic level. Besides if in culture condition the contamination will be less. In addition age is to be another factor to accumulation.

Therefore, it could not be said that all fishes of higher lipid contents contained higher residues level or all fishes of higher trophic levels contained higher. Both the lipid contents and (or) the trophic position were the predictor of concentraitions of organohalogen residues, also reported in the fishes from subarctic lakes in Yukon Terrritoy (kidd *et al*. 1998, 2000). Generally fatty fishes of higher trophic levels

contained higher DDTs residues than the lean fishes of lower trophic levels reported from the fishes from Qiantang River in East China (Kong *et al*. 2005, Zhou *et al*. 1999, 2007).

4.4. Grouping of fishes on the basis of total DDTs concentration using dendogram

To have a clear idea about the grouping of fishes and prawn on the basis of DDTs concentrations, dendogram was constructed using the raw data of total DDTs residues. The two dendograms, first one classifies the fishes and prawn that commonly found in all seasons according to their average annual DDTs accumulations. While the second dendogram classifies the fishes and prawn of summer season only according to their total DDTs accumulations of the season. The second dendogram was conducted on summer season because of the species contained maximum DDTs concentrations in this season.

The first dendogram grouped species into three big clusters with subgroups. The first cluster group contains Kachki, Chanda, Ghainna, Shing in one subgroup and Bata, Sharpunti, Jatpunti, Golda-chingri and Rui in other subgrpup that shows the lower DDTs accumulations. Therefore this Cluster includes all herbivores (Rui, Ghainna and Bata that occupy second trophic level just after the producer), five omnivores (Kachki, Shing, Sharpunti, Jatpunti and Goldachingri that contain low lipid conternt and occupy lower trophic levels as feed mainly on plant materials, planktons etc), one carnivore fish (Chanda that contains low lipid content and occupy lower trophic level as feed mainly on zooplankton). The second Cluster includes Bajaritengra, Chewa in one subgroup and Tengra, Bele, Gulsha, Boal and Foli in other subgroup that shows higher DDTs accumulation. Therefore, this Cluster containes three omnivores (Bajaritengra, Tengra and Gulsha that have higher lipid contents and occupying higher trophic levels that feed mainly on animals; insects, small fishes etc.) and four carnivores (Chewa, Bele, Boal and Foli which have much lower lipid contents and occupy higher trophic level that feed small fishes, mollusks, insects and crustaceans etc.). The Cluster 3 group includes the species of four carnivore fishes Boro baim, Meni, Poa and Bacha that shows much higher DDTs accumulations. In this case

Borobaim mainly feed on bottom sediment, Meni and Poa are highly carnivorous and predatory in nature and Bacha is also highly carnivorous, voracious and containe higher lipid content than the others.

The second dendogram grouped species into four big Clusters with subgroups which is more or less similar to the first one. The Cluster 1group shows the lower DDTs accumulation, the Cluster 2 and 3groups show higher while the Cluster 4 shows much higher DDT accumulations. The Cluster 1group contains just like the first dendogram, all herbivores, five omnivores that have low lipid conternt and occupy lower trophic levels, one carnivore fish that contains low lipid and occupy lower trophic level as feed mainly on zooplankton. The Cluster 2 group includes three omnivores that have higher lipid content and and occupying higher trophic levels that feed mainly on animals; insects, small fishes etc. and three carnivores that have much lower lipid contents and occupy higher trophic level that feed small fishes, mollusks, insects and crustaceans etc. The Cluster 3 group includes the species of three carnivore fishes that are predatory in nature and shows higher DDTs accumulations. The Cluster 4 group contains four carnivores in which case one is bottom sediment feeder, two highly carnivorous and predatory in nature and one is also highly carnivorous, voracious and contained much higher lipid content.

From the cluster analysis it is easy to identify the groups of fishes with lower, medium and higher DDTs residues. Therefore from above discussion, it could be said that lower residues may be found in herbivores, lean and plant based omnivores and lean and zooplankton based carnivores; medium or higher residues may be found in fatty and animal based omnivores, lean and lower carnivores while the fishes with much higher residues may include bottom feeder carnivores, predators and fatty top carnivores.

From the dendogram, it is clear that lipid contents, trophic positions and feeding habit all together influence the accumulation of pesticide residues in fishes.

4.5. DDTs residues and different feeding mode through Principal Component Analysis (PCA)

From the principal component analysis (PCA), all metabolites of DDTs were strongly correlated with the first axis, suggested agricultural run-off as an important source of pollution in the study area. Three distinct groups are formed on the basis of trophic levels of fish species (herbivore, omnivore and carnivore). The maximum concentrations of DDTs were recorded in group carnivores that were relatively higher than omnivores and carnivores. Trophic positions and lipid contents are reliable predictors of organohalogen concentrations in aquatic ecosystem (Kidd *et al.* 2000). DDTs can biomagnify and bioaccumulate in carnivore fish as they are generally positioned at the top of trophic level (Deribe *et al.* 2011, Zhou *et al.* 2007). A significant high concentrations of DDTs in muscle of carnivore fish species has been reported in contrast to other fresh water fish species with various feeding habits by Robinson *et al*. (2016) and Kong *et al*. (2005). Therefore it has been suggested that bioaccumulation of OHs is species-specific in fish due to their ecological features e.g. habitat and feeding modes.

4.6. Meteorological parameters that influence DDTs concentrations

Temperature and humidity affect the toxic potential of environmental pollutants like DDT, has been suggested by Baetjer (1968). In the present study, DDT and its residues showed significant positive correlation with ambient temperature where highest concentrations of DDTs residues in species observed in summer season that had higher ambient temperature and lowest concentrations observed in winter that had lowest temperature. The results were such that higher ambient temperatures enhance the percutaneous absorption of chemicals, has been reported by Suskind (1977) and also the bioconcentration factor or amount accumulated has been shown to increase with temperature for DDT (Cember *et al.* 1978, Boudou *et al.* 1980, Reinert *et al.* 1974). Similar findings also reported by Phillips (1980) that the solubility of several organochlorine insecticides in water increases with ambient water temperature, leading to greater uptake and possibly higher toxicity to aquatic biota and study with DDT and mosquito fish, and rainbow trout support this.

The other parameter, humidy showed significant negetive correlation with DDTs residues that is concentratons of DDTs residues increases with decresing humidity. That might be due to the volatile nature of DDT. As the lower humidity exerts the evaporation rate to higher and which increases DDT burden in the air and finally atmospheric diposition of DDT into aquatic habitat. Baetjer (1968) reported the effect of humidity on toxic chemical.

Rainfall is the additional important parameter to influence the DDTs accumulations. In the present study, rainfall showed negative correlation but not significant effective. The highest bulk rainfall occurred in rainy-season showed lower concentrations of residues. This might be due to dilution effect of contaminants in the aquatic body. But the lowest rainfall in winter season did not show highest concentration of residues because other physiological and biological factors influence together on DDTs accumulations.

In the environment the meteorological factors persist together. Therefore it could not expect that high temperature cause higher accumulation or higher humidity and rainfall cause lower accumulation in all time. In the present study such condition also shown in average highest temperature was in rainy-season but contained lower concerntrations of DDTs and lowest rainfall in winter but contained lower residues level as not expected. Therefore it could be said that temperature, humidity and rainfall have the combined effects on DDTs accumulation. In addition, for instance low metabolism in fishes in cold weather reduce the accumulation of toxic chemicals.

4.7. Seasonal distribution of DDTs

Concentration of total DDTs in fish and prawn species varied greately with the seasons. The species in summer season showed the highest mean concentration and in winter showed lowest concentration where the chronology is summer $>$ autumn $>$ rainy-season > winter and the ANOVA results confirmed that there was indeed a significant differences between seasons (LSD test, $p \leq 0.01$) showed that the concentrations of summer seasons varied significantly with other three seasons. Similar results were observed by the Ntow (2005) that season had large and significant effects on the varience for concentrations of DDTs and fishes accumulated greater pesticide residues during summer season. In the present study, the alarming condition is that during summer season the mean concentration of total DDTs was significantly higher than other three seasons which was around ten times greater than autumn and around twenty times greater than rainy-season and winter. Several factors and conditions may be responsible for this alarming result.

First factor is food intake. Food intake in one of the main route by which fish can accumulate POPs in their natural aquatic habitat (Holden *et. al.* 1996). Fish shows high rate of feeding intensity during summer or premonsoon month may be due to extra energy requirement for building up of gonads (Abbas 2010, Chaklader *et. al.* 2014, Khan *et al.* 1988, Serajuddin *et al.* 1988). Moreover the higher levels of DDTs in fish generally may be ascribed to increase feeding in insect larvae, coarse vegetable matters and sediments associated particles that had accumulated the pesticides over time. As before mentioned during summer and premonsoon the river water shows highest concentration of DDTs and the concentration of DDTs may be several thousand times grater in fish body than the surrounding water because of its bioconcentration and bioaccumulation nature (ATSDR, 2002). So it is clear that the highest concentration of DDTs in river water and highest feeding intensity together with may be responsible for much higher concentration of DDTs in fish tissue during summer.

Second factor is the rainfall. After winter or dry season, the short rainfall of summer and premonsoon season may washout DDTs from the adjacent field to the river that was previously exposed. As the water level was comperatively lower during premonsoon, the runoff DDTs would make the river water contaminated with highest concentration in that season. Similar result was reported for the waterbody of river Sio, river Nzoi and Lake Victoria of Kenya. Where DDT concentration in water of these rivers and lake were highest in short rain season than heavy rain and dry season (Wangdiga and Jumba 2002)

Third is temperature. Rate of residue accumulation and bioconcentration in fish increases with temperature (Kidwell *et al.* 1990). The highest ambient temperature reported in premonsoon may also increase higher DDTs accumulation in fish body during this season. Upadhi and Wokoma in 2012 repoated the same findings that high temperatue increased the higher pesticide residues accumulations in fish tissue in Elechi Creek, Nigeria Delta, Nigeria.

The fourth and well known factor is lipid contents. The highest mean lipid content of fishes reported in premonsoon prior to spawning also reported similar findings in some fishes by Vollenweider *et al.* 2011. Higher lipid content may also responsible for higher DDTs accumulation in the season.

From the above discussion, it is clear that rainfall, food intake, ambient temperature and lipid contents may influence in DDT accumulation. The combined effects that increase its concentration at cumulative rate in fish body. Therefore the mean DDTs concentration of fish and prawn species was much higher in summer season than other seasons. The result of ANOVA with Tukey HSD and LSD statistical test also confirmed that the mean DDTs concentration of fish and prawn species in summer season showed significant differences with other three seasons.

Higher residue levels also detected during autumn compare to rainy and winter season may be attributed to high feeding intensity and concentrated pollutants due to small volume of water in riverbases. During autumn, the fish takes food highly for recovery energy that they loss through spawning and also food availability is much in this season. Therefore higher feeding intensity in autumn may be responsible for higher DDTs accumulation. On the otherhand, as there are two main routes by which fish can bioaccumulate POPs in their natural aquatic habitat: from water via gill and body surface and from the diet (Burreau *et al*. 2004, Campbell *et al*. 2000 and Holden *et al*. 1966). So much pollutant could be entered in the fish body from the surrounding water with high concentration.

On the other hand low residue levels detected during rainy-season compare to the summer and autumn. This could be attributed to dilution effects based on large volumes of rainwater. A similar finding, lower concentration of persitant organis pollutents was also observed in the Lake Victoria of Kenya during wet season (Wandiga and Jamba, 2002). However fresh water fishes spawn mostly during rainyseason, July-september (Rahman 2005, Shafi and Quddus 1982). Organohalogen compounds are transferred from muscle tissue to the gonads along with the fat during the final stage of gonad maturation (Frantzen *et al.* 2011). Therefore, through spawning DDTs may be decreased in fish body during rainy-season. Maternal transfer of organohalogen compounds to offspring via gonads has been shown for oviparous fish species such as zebrafish and salmonoids (Russell *et al.* 1999, Heiden *et al.* 2005, Nyholm *et al*. 2008). Moreover feeding intensity of fishes is lower during spawing season (Jhingaran 1961, Desai 1970, Bhatnagar and Karamchandani 1979, Fatima and Khan 1991) may be related to lower DDTs accumulation

Lowest residues levels also detected during winter could be attributed by the lower feeding intensity (Abbas 2010), lower lipid content and also lower ambient temperature. During winter season, the feeding intensity of fishes is very much low due to scarcity of food, moreover they loss of stored lipid contents through metabolism to supply energy for physiological activities. In the season the ambient temperature was lowest may also be related to lower DDTs accumulation.

4.8. Comparison with previous work

Comparison with results from other water bodies of Bangladesh is difficult because all the relevant studies have been done with species collected from different ecological habitat, small sizes of the samples and also samples collected in a single season or randomly from market etc. Furthermore, the results found are expressed in different units. In a recent study (Hossain *et al.* 2016) on twenty two different fish species of Kangsha and Titas river, the concentrations of DDTs in several fish species were similar to concentrations that were found in the present study while the range of concentration (4.71-78.81 ng g^{-1}) was similar to range of rainy and winter season's samples of the present study. The concentration levels of DDTs (0.02-1.4 mg/kg or

20-1400 ng g^{-1}) reported by Nahar *et al.* (2008), were also closely similar to the present study. On the otherhand, most of the rearch on DDTs concentration in our country, were held on dry fishes for the direct and excessive use of this contaminant in dry fish where the concentration levels were higher to the present study reported by Bhuiyan *et al*. (2009), Hasan *et al*. (2013) and Hasan *et al.* (2014)

The results of the present investigation are comparable with other studies done in the world. For example, Cal *et al.* 2008 found that the levels of DDT and its metabolites residues in fishes from Cinca river, a tributary of Ebro river of Spain were higher (nondetected to 2098 μ g kg⁻¹ or ng g⁻¹) than the residue levels that analysed in the present study. Concentration levels of DDTs residues in fishes of Midcontinental great river (US), alpine lake (Switzerland), rocky mountains (Canada), Tibetan plateau (Central Asia) were reported in lower levels (Blocksom *et al.* 2010, Schemid *et al*. 2007, Demers *et al.* 2007 and Yang *et al.* 2010) than the present study.

As mentioned above, in most of the studies where concentration levels analysed only for one or two seasons while the present study includes four seasons as in, our country four seasons are distinct and the meteorological, physical, biological and many other factors vary with these seasons which direcly or idirectly affects the residue levels. Therefore it could be said that the present study shows the actual status of contaminant levels in fishes throughout the year which is quite difficult to compare with other study with limited data.

4.9. Detection of DDT exposure time from the ratio of (DDE+DDD)/DDT

Metabolites of some organochlorine pesticides such as DDT have different concentration ratios in the environment, thereby indicating different contamination sources. Specific ratios of parent and metabolites of organochlorine pesticides compounds have been widely used to identify past and present input application into the environment (Walker *et al.* 1999).

The ratio of (DDE+DDD)/DDT is a helpful tool in revaling the significance of the degradation of DDT and to evaluate the current or past use of DDT in the region (Yu *et al.* 2011). The ratio higher than 0.5 indicates past input of DDT while lower indicates recent input of DDT. In the present study, 38.46% samples having value of (DDE+DDD)/DDT lower than 0.5 ratios. These result indicated that the presence of new DDT inputs in the environments of Meghna river and flood plains of Sonargaon area. Fresh inputs of DDTs were also reported in India (Chourasiya *et al.* 2014) Uganda (Sseburgene *et. al*. 2009), South China (Hao *et al*. 2014, Sun *et al*. 2014, Zhao *et al.* 2009) and in Brazil (Rissato *et al.* 2007) etc.

Moreover tropical data on POPs shows rapid degradation patterns, low residual levels and wide distribution because the tropics are characterized by different radiation throughout the year, a high load of microorganisms, tropical rain and various soil types(Wandiga *et al*. 2001, Peters *et al.* 2001).

4.10. Tolerance limits

Food consumption is the main exposure route for organochlorinated pollutants for the general population, and fish and fishery products seem to be the main contributors to the total dietary consumption of these pollutants (Schnitzler *et al.* 2011). Therefore, fish consumption can be a major source of human exposure to OCPs in Bangladesh, and the exposure levels depend on the lipid content of the fish and the amount of fish consumed.

No regulations by Bangladesh agriculture or health agencies are available regarding maximum OCP limits in fish and other food. To roughly evaluate the potential environmental risk of organochlorine residues, concentrations of selected analytes in tissues of the present study were compared in with the maximum residue limits (MRLs) recommended for human consumption by various agencies.

[Maximum Residue Level](http://www.efsa.europa.eu/en/pesticides/mrls.htm) (MRL) is the highest level of a pesticide residue that is legally tolerated in or on food or feed when pesticides are applied correctly. Good Agricultural Practice will ensure a concern safety to human health. The concentration of organochlorine pesticides obtained in various food samples were compared with MRLs, set by the various international agencies such as Food and Agricultural
Organisation/World Health Organisation (FAO/WHO 2012). The values found in fish tissue in four different seasons are below the MRL value of 5000 ng g^{-1} w.w. for consumption set by Codex Alimentarius Commission of FAO/WHO (2012). On the other hand, the European Union (Binelli and Provini, 2004) has established a maximum admissible limit of 50 ng g^{-1} ww for DDTs, in the present study, 20.83% species of rainy-season, 68.18% of autumn, 13.04% in winter and 100% in summer exceeded this limit. Moreover, European Union Directive 1999/788 has established the maximum acceptable limit of 200 ng g^{-1} ww for human consumption (Robinson *et*) *al.*2016). Only 4.55% species of autumn and 86.36% of summer samples exceeded this limit. However, 4.55% species of autumn and 36.36% of summer are above concentrations associated with reproductive toxicity in several species of fish (500 ng g^{-1} w.w.; Jarvinen and Ankley 1999).

4.11. Human Health risk estimates

From the obtained concentrations of DDTs residues, dietary exposure and health risks were calculated for adult. The estimated daily intake and Health Risk Index (HI) were calculated for the contaminant of each fishes. The HI showed that total DDTs residues in all fishes recorded less than 1 (HI<1).

The statistic of daily food intake of an adult person of our country reveals that fish rank fourth after rice, cereal and vegetables which contribute about 4% of the total daily food consumption in weight (HIES 2010). Although detail study was lacking about the DDTs residues in our food items, some previous works reported the detection of the contaminants in some food such as chicken, dry fish and vegetables (Nahar *et al.* 2008). Therefore it can be said that although the HI of the studied fishes are less than 1 but toghther with other food items of daily meal, total HI would be greater than 1. This shows that there is health risk associoated with lifetime consumption of the studied fishes, togther with other contaminated foods.

Similar findings also reported by the Nuapia *et al.* (2016) that the values of HI of fishes were <1 but together with vegetables and meat the values were >1 and that associated with human health hazards.

5. SUMMARY, CONCLUSION AND RECOMMENDATIONS

5.1. SUMMERY

Pesticide residue problem is an environmental hazard and becoming serious focus for human health. Organohalogen pesticides such as DDT and its metabolites are of great concern to the environmental scientists for several decades, due to their persistence, bioaccumulation, long-range transport, toxicity and adverse effects on environment and human health including reproduction and birth defects, immune system dysfunction, endocrine disruptions and canceR. The intensive cultivation of the agriproducts depends upon the use of fertilizer, pesticides, insecticides, fungicides and herbicides. About 25% of these compounds pass to the nearby water-body and act as a pollutant sources for fish and other aquatic organisms. Fishes are used extensively for environmental monitoring because they uptake contaminants directly from water and food. Generally the ability of the fish to metabolize organohalogen is moderate; therefore, contaminants load in fish are well reflective of the state of pollution in surrounding environment (Matin *et al.* 1998). Thus in view of multidimensional impacts of pesticide in fish, environment and human, the scope of the present study is to asses some organohalogen pesticide residues in fishes of Bangladesh. In this study, the concentrations of organohalogen pesticide residues; DDT and its metabolites DDE, DDD, 2,4-DDT and 4,4-DDT were investigated in different fish and prawn species of Meghna river

Different fish and prawn species of different trophic levels were collected from Meghna river at Boidyer Bazar of Sonargaon Upazila of Narayanganj District. Samples were collected periodically from four different seasons; rainy-season, autum, winter and summer. The samples were immediately wrapped with aluminium foil, put into a chill box with ice and transported to the laboratory on the same day. After identification and taking some physiological data, the samples were then stored in a freezer at a temperature below -20°C temperature until analysis. Before extraction, the fish tissue was made bone free, chopped and blended. The samples were extracted following QuEChERS method using ethyle acetate as solvent and cleaned up using sulphuric acid. The cleaned extracts were analyzed by GC-ECD for the residual amounts of DDTs (DDT and its metabolites). The calibration curve of serially diluted certified DDTs (DDE,DDD, 2,4' DDT and 4,4' DDT) standard solution showed peak area with a very linear correlation (r^2 =0.996, r^2 =0.995, r^2 =0.985 and r^2 =0.996 respectively). LOD and LOQ were found to be 0.0625 ng g^{-1} , 6.25 ng g^{-1} respectively. The percent of the recovery DDTs in spiked control Rui fish samples at different levels (0.05, 0.10 and 0.20 ng g^{-1}) The percentage recoveries for fish samples were found to be 88-92 % for DDE, 101-113% for DDD, 76-104 % for 2,4'-DDT and 70-90 % for 4,4'-DDT.

5.1.1. Lipid contents of fish and prawn samples of different seasons

The lipid contents (%) of different fish and prawn species ranged from $0.53 \pm 0.01\%$ in Boal to $13.98 \pm 1.50\%$ in Bacha during rainy season, $0.51 \pm 0.06\%$ in Boal to 14.43 \pm 0.64% in Rui during autumn, 0.41 \pm 0.07 in Shole to 13.66 \pm 0.56% in Rui during winter and $0.12 \pm 0.01\%$ in Boal to 15.40 \pm 0.01 % in Bacha during summer. Variation of lipid contents is species specific. As DDT accumulates in the fat body of organisms due to its lipophilic nature, lipid contents of the fishes also studied due to analyse their relationsphip.

5.1.2. Total DDTs residues of fishes and prawn of different seasons

The mean concentrations of Σ DDTs residue (ng g⁻¹ ww) in fishes and prawn during rainy-season ranged from 2.64 \pm 0.35 ng g⁻¹in Kachki to 191.14 \pm 31.18 ng g⁻¹ in Bacha, 16.42 ± 1.90 ng g⁻¹ in Kalisha to 271.50 \pm 6.17 ng g⁻¹ in Bacha during autumn, 3.88 \pm 0.60 ng g⁻¹ in Golda Chingri to 141.57 \pm 10.24 ng g⁻¹ in Bacha during winter and 157.58 \pm 1.15 ng g⁻¹ in Kachki to 1660.89 \pm 157.9 ng g⁻¹ in Bacha during summer. Considering the mean concentrations of DDTs residue of four seasons, Rui, Ghainna, Bata, Jatpunti, Sharpunti, Shing, Magur, Kachki, Goldachingri, Gutum, Khalisha, Kaikka, Tarabaim and Boal conained lower amount of DDTs residue While Tengra, Bajari-tengra, Gulsha, Borobaim, Chewa, Meni, Bele, Poa and Bacha contained higher amont of DDTs residues.

The mean concentration of total DDTs for all the herbivores, omnivores and carnivores were 10.38, 21.38 and 45.10 ng g^{-1} during rainy-season, 36.37, 47.34, 96.00 ng g^{-1} .during autumn, 16.75, 20.32 and 43.30 ng g^{-1} .during winter and 264.64, 334.71, 845.83 ng g^{-1} during summer. Therfore in every season on the basis of average total DDTs residues the chronlolgy was herbivore < omnivore < carnivore. One way ANOVA with Tukey HSD and LSD Post Hoc tests showed significant differences in total DDTs between herbivore, omnivore and carnivore fishes.

In the present study, fishes with lower DDTs residues may be attributed to their lower position in food chain or low lipid content and sometimes both of these. As mentioned above, Rui, Ghainna and Bata contained lower DDTs residues may be related to their lower position in foodchain that feed on plant materials while the lower residue levels of Kachki, Chanda, Shing fish and Goldachingri may be related to their lower lipid content and positioned in lower trophic levels that mainly feed on zooplankton and plant materials.

Jatpunti and Sharpunti contained comperatively lower DDTs residues but higher than Kachki, Shing , Magur, Chanda, Rui, Ghainna and Bata may be attributed to their higher lipid content and feeding habit also that mainly feed on algae, plant material and small amount insects due to omnivorous nature

Foli and Boal also contained lower amount of DDTs residues. Although they occupy in the higher trophic level but contained low amount of lipid content may be responsible for lower amount of DDTs residues.

Tengra, Bajari-tengra and Gulsha contained higher DDTs residues may be related to their both higher lipid content and higher trophic positions as feed mainly on small fishes, insects, moluscks and little amount of algae and plant material due to their omnivorous nature.

Chewa, Meni, Bele, Poa,Shole, Gojar and Bacha contained much higher amount of DDTs residues. All these fishes are highly carnivorous,voracious, predatory in nature,

and top consumer as feed on small fish, juvenile of large fish, frog, snake, insects, earthworm and tadpole larva etc. due to their highly carnivorous and predatory in nature, which could be related to their higher DDTs accumulation. While Borobaim contained higher residues due to their bottom dwelling carnivorous nature.

Fishes are classified into different cluster on the basis of DDTs residues through dendogram. From the cluster analysis, it is clear that herbivore fishes, lean and plant based omnivores and lean, zooplankton based carnivores belong to same group, contained lower DDTs residues; fatty and animal based omnivores, lean and lower carnivores contained medium or higher DDTs residues while bottom feeder carnivores, predators and fatty top carnivore fishes belong to same group that contained much higher residues.

Therefore, it is clear that lipid contents, trophic positions and feeding habit all together influence the accumulation of pesticide residues in fishes.

5.1.3. Meteorological parameters and DDTs concentrations

In the present study together with biological factors the meteorological parameters (temperature, humidity and rainfall) are also considered to asses the accumulation of DDTs in fish body. From the analysis of distribution pattern of contaminats in fish and prawn tissues, it is clear that both biological and analytical factors are important to interpret DDTs bioaccumulation.

In the present study, DDTs residues showed significant positive correlation with ambient temperature and significant negative correlation with humidity. Rainfall is other important parameter to influence the DDTs accumulations showed negative correlation but not significant.

Therefore all these physical factors are also important to predict the DDTs level but they donot work singlely as persist together in the environment and influence together to accumulation DDTs.

5.1.4. Seasonal distribution of DDTs concentrations

Lipid contents and ∑DDTs concentraitons vary significantly between species for all seasons. One way ANOVA with Tukey HSD and LSD Post Hoc tests showed highly significant differences in total DDTs (∑DDTs) concentrations of fishes and prawn between seasons $(p<0.01)$ where the summer season showed highly significant variations between rainy-season, autumn and winter. The biological and meteorological factors that influence DDTs accumulation, vary greately with seasons may be attributed to the significant seasonal variation of residues.

In case of ∑DDTs concentrations 70% species showed the highest amount in summer and lowest in rainy-season. Considering the mean values DDTs residues for seasons the chronology is winter $(25.60 \text{ ng/g}) < \text{rainy-season}$ $(30.88 \text{ ng/g}) < \text{autumn}$ (63.34 m) ng/g < summer (580.72 ng/g).

During summer season, the mean concentration of ∑DDTs was significantly higher than other three seasons which was around ten times greater than autumn and around twenty times greater than rainy-season and winter. Several factors and conditions influence together to higher accumulation. The factors are higher food intake, high ambient temperature and lipid contents , short rain in summer may influence together in DDT accumulation that increase its concentration at cumulative rate in fish body.

Higher residue levels also detected during autumn compare to rainy-season and winter may be attributed to high feeding intensity for recovery energy that they loss through spawning in rainy-season and also food availability is much in this season and concentrated pollutants due to small volume of riverwater.

On the other hand low residue levels detected during rainy-season compate to the summer and autumn. This could be attributed to dilution effects based on large volumes of rainwater. Rainy-season is the spawning season of fishes therefore, fishes transfer DDTs residues to offspring via gonads. Moreover feeding internsity of fishes is lower during spawing season may be related to lower DDTs accumulation

Lowest residues levels also detected during winter could be attributed to lower feeding intensity, lower lipid content and also lower ambient temperature. During winter season, the feeding internsity of fishes is very much low due to scarcity of food, moreover they loss stored lipid contents through metabolism to supply energy for physiological activities. In this season the ambient temperature was lowest may also be related to lower DDTs accumulation.

5.1.5. Tolerance limits

The values found in fish tissue in four different seasons are below the MRL value for consumption set by Codex Alimentarius Commission of FAO/WHO (2012) but exceeded the maximum admissible limit and maximum acceptable limit set by other international agencies. However, in some cases the concentration levels exceeded the limit that may be associated with reproductive toxicity in several species of fish.

5.1.6. Present and historical use of DDT

From the ratio of (DDE+DDD)/DDT, it can be evaluated the current or past use of DDT in the region. In the present study, analyzing the ratios indicates the evidence of both current and historical use of DDT as pesticides in the neighbouring environments of Meghna river and flood plains of Sonargaon area. Therefore, it can be said that DDT is still exposed to the environments in the country.

5.1.7. Human health risk

As the residue levels exceeded some tolerance limits together with their continuous exposure make these contaminants as major concer to human health hazard. From this point of view the human heath risk was estimated for the studied fish species.

Hazard Index (HI) is used to assess the health risk of consumers from the intake of pesticides contaminated fish. The estimated HIs is obtained by dividing the Estimated Daily Intake (EDI) by their corresponding values of Acceptable Daily Intakes (ADI) by WHO/FAO (FAO/WHO 2010). If the value of HI is greater (>) than 1 that would be associated to health risk.

All the detected HIs of DDTs residues for studied fish and prawn species were <1. Though the HIs are less than 1 but toghther with other food items of daily meal, total HI would be greater than 1. This shows that there is health risk associoated with lifetime consumption of the studied fishes, togther with other contaminated foods.

5.2. CONCLUSION

Organohalogen compounds especially dichlorodiphenyltrichloroethanes (DDTs) become a worldwide concern, despite controls on their manufacturing, importation and agricultural or vector practices. Through their lipophilicity and persistence, the toxic chemicals and their residues may concentrate in the fatty tissues of animals leading to environmental persistence, bioconcentration and biomagnificaiton through the food chain. Because these chemicals are toxic to living organisms, increased accumulation in the food chain may pose serious health hazards to human and wildlife.

From this comprehensive analysis of DDT and its metabolites (4,4²DDT, 2,4²) DDT,DDD and DDE) in reverine fishes and prawn species of different feeding habits at different seasons, demonstrating that the meteorological and biological parameters influencing the contamination, lead to the environmental risk of fish species for human health issues. The widespread year round occurance of DDTs in all fishes and prawn species confirms the recent use of this illegal compound of potential contaminant to fish and environment. Moreover the significant levels of DDT metabolites indicating the past use of the parent compound (DDT).

The levels of total DDTs residues in most cases exceeded the prescribed maximum admissible and acceptable limits set by different international agencies; European Union and FAO/WHO, though below the limit of [Maximum Residue Level](http://www.efsa.europa.eu/en/pesticides/mrls.htm) (MRL) for consumption set by Codex Alimentarius Commission of FAO/WHO (2012). It is predicted that these residues may be originated from present and also past activities in agricultural practices. The continuous consumption of contaminated fishes throughout the year with significant amount of the toxic eliments may pose human health hazards. From the estimated Hazarad Index (HI) of studied fishes, it can be said that the fishes together with other contaminations in food further cause complex hazard issues. To prevent this kind of health disaster, it is essential to formulate a regular monitoring practice on the residues of the key components of human food chain. Further policy forming and awareness building on this issue is of prime need.

5.3. RECOMMENDATIONS

- Since fish is common and essential part of our diet, this study showed serious concern on food born pesticide contaminant like illegal DDT in fish tissue
- The control of illegal agrochemicals is must, as well the strong monitoring and legal control is necessary
- More understanding on agrochemicals instruction in food chart and the food safety issues is essential
- As DDT is not produced locally, stopping illegal entry of these long persistent substances in the country is of great importance for human health issues.

LITERATURE CITED

- ABBAS, A. 2010. Food and feeding habits of freshwater catfish, *Eutropiichthys vacha. Indian J. Sci. Res.***1(2)**: 83-86.
- ACKERMAN, L.K., SCHWINDT, A.R., MASSEY SIMONICH, S.L., KOCH, D.C., BLETT, T.F. and SCHRECK, C.B. 2008. Atmospherically deposited PBDEs, pesticides, PCBs, and PAHs in Western US National Park fish: concentrations and consumption guidelines. *Environ. Sci. Technol.* **42**: 2334–2341.
- ADEYEMI, D., UKOKPO, G., ANYAKORA, C. and UNYIMADU, J.P. 2008. Organochlorine pesticide residues in food samples from Lagos markets, Nigeria. *Am. J. Environ. Sci.* **4 (6)**: 649-653.
- ADU-KUMI, S., KAWANO, M., SHIKI, Y., YEBOAH, P., CARBOO, D. and PWAMANG, J. 2010. Organochlorine pesticides (OCPs), dioxin-like polychlorinated biphenyls (dl-PCBs), polychlorinated dibenzo-p-dioxins and polychlorinated dibenzo furans (PCDD/Fs) in edible fish from Lake Volta, Lake Bosumtwi and Weija Lake in Ghana. *Chemosphere* **81**: 675–684.
- AKOTO, O., ASORE, A., AZUURE and ADOTEY, K.D. 2016. Pesticide residues in water, sediment and fish from Tono Reservoir and their health risk implications. *SpringerPlus* **5**: 1849
- ALAMDAR, A., SYED, J.H., MALIK, RN, KASTSOYIANNIS, A., LIU, J. and LI, J. 2014. Organochlorine pesticides in surface soils from obsolete pesticide dumping ground in Hyderabad City, Pakistan: contamination levels and their potential for air–soil exchange. *Sci. Total Environ*. **470:** 733–741.
- ALI, N, SAMAS, E., MALIK, R.N., NEELS, H. and COVACI, A. 2013a. Organohalogenated contaminants (OHCs) in human serum of mothers and children from Pakistan with urban and rural residential settings. *Sci. Total. Environ.***461**: 655–662.
- ALI, N., MALIK, R.N., MEHEDI, T., SAMAS, E., JAVEED, A. and NEELS, H. 2013b. Organohalogenated contaminants (OHCs) in the serum and hair of pet cats and dogs: biosentinels of indoor pollution. *Sci. Total Environ*. **449**: 29–36.
- AMARANENI, S.R. 2006. Distribution of pesticides, PAHs and heavy metals in prawn ponds near Kolleru lake wetland, India. *Environ. Int*. **32**: 294–302.
- AMIN, S.M.N., ARA, R., MOHAMMAD, H. and ARSHAD, A. 2014. Food habits of snakehead, *Channa striatus* (Bloch), in the lotic streams of Universiti Putra Malaysia, Malaysia. *J. Food, Agricul. Environ.* **12(2)***:* 979 -981.
- AMODIO, E., TURCI, R., MASSENTI, M.F., DI GAUDIO, F., MINOIA, C.and VITALE, F. 2012. Serum concentrations of persistent organic pollutants (POPs) in the inhabitants of a Sicilian city. *Chemosphere* **89**: 970–974.
- AONO, S., TANABE, S., FUDISE, Y., KATO, H. and Tatsukawa, R. 1997). Persistent organochlorines in minke whole (*balaenoptera acutrostrata*) and their prey species from the Antarctic and the North Pacific. *Environ. Pollut.* **98**: 81-89.
- AQUASTAT. 2011. Survey report on Ganges-Brahmaputra-Meghna River basin. FAO.
- ARNOT, J.A. and GOBAS, F.A.P.C. 2004. Food web bioaccumulation model for organic chemicals in aquatic ecosystems. *Environ. Toxicol*. *Chem.* **23:** 2343- 2355.
- ATSDR, 2002. Toxicological Profile for DDT, DDE, DDD. US Department of Health and Human Services, Public Health Service, Atlanta, GA. http://www.atsdr.cdc.gov/toxprofiles/tp35-p.pdf.
- AYEJUYO, O.O., WILLIAMS, A.B. and IGBASAN, S.O. 2008. Assessment of organochlorine pesticide residues in irrigation groundwater of Lagos. *J. Chem. Soc. Niger*. **33:** 65-69.
- BAETJER, A. M. 1968. Role of environmental temperature and humidity in susceptibility to disease. *Archs. Environ. Hum.* **16:** 565-570.
- BARAKAT, A.O., KIM, M., QIAN, Y. and WADE, T.L. 2002. Organochlorine pesticides and PCB residues in sediments of Alexandria harbour, Egypt. *Mar. Pollut. Bull*. **44**:1421–1434.
- BARRIADA-PEREIRA, M., GONZALEZ-CASTRO, M.J., MUNIATEGUI-LORENZO, S., LOPEZ-MAHIA, P., PRADA-DODRIGUEZ, D. and FERNANDEZ-FERNANDEZ, E. 2005. Organochlorine pesticides

accumulation and degradation products in vegetation samples of a contaminated area in Galicia (NWSpain). *Chemosphere* **58**: 1571–1578.

- BEYER, A., MACKAY, D., MATTHIES, M., WANIA, F. and WEBSTER, E. 2000. Assessing Long-Range Transport Potential of Persistent Organic Pollutants. *Environ. Sci. Technol.* **34(4)**: 699–703.
- BELTON, B. and Little, D.C. 2011. Immanent and interventionist inland Asian aquaculture development and its outcomes. *Develop. Policy Rev.* **29 (4):** 459– 484.
- BES (Bangladesh Economic Survey) 2014. A report prepared under the programme, Independent Review of Bangladesh's development (IRBD), CPD, released on 2014.
- BHANTI, M. and TANEJA, A. 2005. Monitoring of organochlorine pesticide residues in summer and winter vegetables from Agra, India.– A case study. *Environ. Monitor Assess.* **110**: 341–346.
- BHATNAGAR, G.K. and KARAMCHANDANI, S.J. 1970. Food and feeding habit of *Labeo fimbriatus* (Bloch) in river Narbada near Hosangabad (M.P.). *J. Inland Fish. Soc. India.* **2**: 30-50.
- BHATTACHARYA, B., SARKAR, S.K. and MUKHERJEE, N. 2003. Organochlorine pesticide residues in sediments of a tropical mangrove estuary, India: implications for monitoring. *Environ. Int*. **29**: 587–92.
- BHUIYAN, M.N.H. 2008. Screening of Organochlorine Insecticides (DDT and Heptachlor) in Dry Fish Available in Bangladesh. *Bangladesh J. Pharmacol.* **3:** 114.
- BHUIYAN, M.N.H., BHUIYAN, H.R., NATH, K.K., AHMED, K., HASSAN, M.T. and BHUIYAN, M.N.I. 2009. Organochlorine Insecticides (DDT and Heptachlore) in dry fish available in Bangladesh: Seasonal trends and species variability. *J. Chil. Chem. Soc.* **54(3)**: 278-281.
- BIDLEMAN, T.F., COTHAM, W.E. and ADDISON, R.F., *et al.* 1992. Organic contaminants in the northwest Atlantic atmosphere at Sable Island, Nova Scotia 1988-1989. *Chemosphere* **24**: 1389-1412.
- BINELLI, and PROVINI, A. 2003. DDT is still a problem in developed countries: the heavy pollution of Lake Maggiore. *Chemosphere* **52**: 717–723.
- BINELLI, A. and PROVINI, A. 2004. Risk for human health of some POPs due to fish from Lake Iseo. *Ecotoxicol. Environ. Saf.* **58**: 139–145.
- BISCOE, L. M., MUTERO, M. C. and KRAMER, A. R. 2004. Current policy and status of DDT use for malaria control in Ethiopia, Uganda, Kenya and South Africa. *Working paper 95.* International Water Management Institute.
- BLOCKSOM, K.A., WALTERS, D.M., JICHA, T.M., LAZORACHAK, J.M., ANGRADI, T.R. and BOLGRIEN, D.W. 2010. Persistent organic pollutants in fish tissue in the mid-continental great rivers of the United States. *Sci. Total Environ.* **408**: 1180–1189.
- BLOCKSOM, K.A., WALTERS, D.M., JICHA, T.M., LAZORCHAK, J.M., ANGRADI, T.R. and BOLGRIEN, D.W. 2010. Persistent organic pollutants in fish tissue in the mid-continental great rivers of the United States. *Sci. Total Environ*. **408**: 1180–1189.
- BORNMAN, M.S, VAN-VUREN, J.H.J., BOUWMAN, H., DE JAGER, C., GENTHE, B. and BARNHOORN, I.E.J. 2007. Endocrine Disruptive Activity and the Potential Health Risk in an Urban Nature Reserve. WRC report No. 1505/1/07. Water Research Commission, Pretoria, South Africa.
- BOUDOU, A., RIBEYRE, F., DELACHRE, A., and MARTY, R. 1980. Bioaccumulation et bioamplification des derives du mercure par un consommateur de troisieme ordre: *Salmo* gairdneri-incidences du facteur temperature. *Wat. Res.* **14:** 61-65.
- BOUWMAN, H, KYLIN, H., SEREDA, B., and BORNMAN, R. 2012. High levels of DDT in breast milk: Intake, risk, lactation duration, and involvement of gender. *Environ. Poll.* **170**: 63-70.
- BOUWMAN, H., COOPPAN, R.M., REINECKE, A.J. and BECKER, P.J. 1990. Levels of DDT and metabolites in breast milk from Kwa-Zulu mothers after DDT application for malaria control. *Bull. WHO.* **88**: 761-768.
- BOUWMAN, H., POLDER, A., VENTER, B., SKAARE, J.U.S., 2008. Organochlorine contaminants in cormorant, darter, egret, and ibis eggs from South Africa. *Chemosphere Bull*. **48**: 30–43.
- BREIVIK, K., ALCOCK, R., Li, Y.F., BAILEY, R.E., FIEDLER, H.and PACYNA, J.M. 2004. Primary sources of selected POPs: regional and global scale emission inventories. *Environ. Pollut.* **128**: 3–16.
- BURREAU, S., ZEBBUHR, Y., BORMAN, D. and ISHAQ, R. 2004. Biomagnification of polychlorinated biphenyle ethers (PBDEs) studies in pike (*Esox lucius*) , perch (*Perca fluviatilis*) and roach (*Rutilus rutilus*) from Baltic Sea. *Chemosphere.* **55**: 1043-52.
- CAL, A.D.L., ELAJARRT, E., RALDU, D., DURA, C. and BARCELO, D. 2008. Spatial variation of DDT and its metabolites in fish and sediment from Cinca River, a tributary of Ebro River (Spain). *Chemosphere* **70**: 1182–1189
- CALDAS, E.D., COELHO, R, LCKR, S. and SIBA, S.C. 1999. Organochlorine pesticides in water, sediment, and fish of Paranoá Lake of Brasilia, Brazil. *Bull. Environ. Contam. Toxicol.* **62**: 199–206
- CAMPBELL, L.M., SCHINDLER, D.W., MUIR, D.C.G., DONALD, D.B. and KIDDA, K.A. 2000. Organochlorine transfer in the food web of subalpine Bow Lake Banff National Park. *Can. J. Fish. Aquat*. *Sci*. **57**: 1258-1269.
- CANABANA, G. and RASMUSSEN, J.B. 1994. Modelling food chain structure and contaminant bioaccumulation using stable nitrogen isotopes. *Nature* **372:** 255- 257.
- CARSON, R. 1962. *Silent Spring*. Boston, New York: Houghton Mifflin.
- CEMBER, H., CURTIS, E. H., and BLAYLOCK, B. G. 1978. Mercury biconcentration in fish: temperature and concentration effects. *Environ. Pollution.* **17:** 311-319.
- CHAKLADER, M.R., NAHAR, A., SIDDIK, M.A.B. and SHARKER, R. 2014. Feeding Habits and Diet Composition of Asian Catfish *Mystus vittatus* (Bloch, 1794) in Shallow Water of an Impacted Coastal Habitat. *World J. Fish. Marine Sci.* **6(6)**: 551-556.
- CHOURASIYA, S., KHILLARE, P.S. and JYETHI, D.S. 2014. Health risk assessment of organochlorine pesticide exposure through dietary intake of vegetables grown in the periurban sites of Delhi, India. *Environ. Sci. Pollut. Res.* DOI 10.1007/s11356-014-3791-x.
- COLBORN, T., and SMOLEN, M.J. 1996. Epidemiological anlysis of persistent organochlorine contaminants in cetaceans. *Rev. Environ. Contain. Toxicol.* **146**: 91-172.
- CONNELL, D.W. 1995. Prediction of bioconcentration and related lethal and sublethal effects with aquatic organisms. *Mar. Pollut Bull.* **31**: 201-205.
- COVACI, A., GHEORGHE, A., HULEA, O.and SCHEPENS, P. 2006. Levels and distribution of organochlorine pesticides, polychlorinated biphenyls and polybrominated diphenyl ethers in sediments and biota from the Danube Delta, Romania. *Environ. Pollut.* **140**: 136–149.
- CROSLY, R.W., DONALD, D.B. and BLOCK, H.O. 1998. Trends and seasonality in a- and g-hexachlorocyclohexane in western Canadian surface waters (1975– 1994). *Environ. Pollut*. **103:** 277–285.
- DA CUNA, R., REY VAZQUEZ, G., PIOL, N., VERREGIA GUERRERO, N., MAGGESE, M.C. and LO, NOSTRO, F. 2011. Assessment of the acute toxicity of the organochlorine pesticide endosulfan in Cichlasoma dimerus (teleostei, perciformes). *Ecotoxicol. Environ. Saf.* **74**: 1065–1073.
- DAHLE, S., SAVINOV, V.M., MATISHOV, G.G., EVENSET, A and NACS, K. 2003. Polycyclic Aromatic Hydrocarbons (PAHs) in Bottom Sediments of the Kara Sea Shelf, Gulf of Ob and Yanisei Bay. *Sci. Total Environ.* **306**: 57-71.
- DALY, G.L. LEI, Y.D., TEIXEIRA, C., MUIR, C.G., CASTILLO, L.E. and WANIA, F. 2007. Accumulation of current-use pesticides in neotropical montane forests. *Environ. Sci. Technol.* **41:** 1118–1123.
- DARKO, G. and ACQUAAH, S. O. 2007. Level of organochlorine pesticides residues in meat. *Int. J. Environ. Sci. Tech*. **4(4)**: 521-524.
- DARKO, G., AKOTO, O.and OPPONG, C. 2008. Persistent organochlorine pesticide residues in fish, sediments and water from Lake Bosomtwi, Ghana. *Chemosphere* **72**: 21–24.
- DAVODI, M., ESMAILI-SARI, A. and BAHRAMIFARR, N. 2011. Concentration of polychlorinated biphenyls and organochlorine pesticides in some edible fish species from the Shadegan Marshes (Iran). *Ecotoxicol. Environ. Saf*. **74:** 294– 300.
- DE MORA, S., VILLENEUVE, J.P., SHEIKHOLESLAMI, M.R., CATTINI, C. and TOLOSA, I. 2001. Organochlorinated compounds in Caspian Sea sediments. *Mar. Pollut. Bull*. **48**: 30–43.
- DEMERS, M.J., KELLY, E.N., BLAIS, J.M., PICK, F.R., St. LOUIS, V.L. and SCHINDLER, D.W., 2007. Organochlorine compounds in trout from lakes over a 1600 meter elevation gradient in the Canadian Rocky Mountains. *Environ. Sci. Technol.* **41**: 2723–2729.
- DERIBE, E., ROSSELAND, B.O., BORGSTROM, R., SALBU, B., GEBREMARIAM, Z. and DADEBO, E. 2011. Bioaccumulation of persistent organic pollutants (POPs) in fish species from Lake Koka, Ethiopia: the influence of lipid content and trophic position. *Sci. Total Environ*. **410**: 136– 145.
- DESAI, V.R. 1970. Studies on the fishery and biology of *Tor tor* (Ham) from river Narbada *J. Indian fish. Soc. India*. **2**: 101-112.
- DEWAILLY, E D. S., VERREAULT, R., AYOTTE, P., SAUVE, L., MORIN, J. and BRISSON, J. 1994. High organochlorine body burden in womenwith estrogen receptor-positive breast cancer. *J. Natl. Cancer Inst*. **86**: 232-234.
- DIETZ, R., BOSSI, R., RIGET, F.F., SONNE, C. and BORN, E.W. 2008. Increasing perfluoroalkyl contaminants in east Greenland polar bears (*Ursus maritinus*): a new toxic threat to the Arctic bears. *Environ. Sci. Technol*. **42**: 2701-2707.
- DOONG, R.A., PENG, C.K., SUN, Y.C. and LIAO, P.L. 2002. Composition and distribution of organochlorine pesticide residues in surface sediments from Wu-Shi River Estuary, Taiwan. *Mar. Pollut. Bull.* **45**: 246-253.
- DUA, V., KUMARI, R., JOHRI, R., OJHA, V., SHUKLA, R. and SHARMA, V. 1998. Organochlorine insecticide residues in water from five lakes of Nainital (UP), India. *Bull. Environ. Contam. Toxicol*. **60**: 209–15.
- EDWARDS, C. A . 1987. The environmental impact of pesticides. *Parasites* **86:** 309- 329.
- El NEMR, A. 2011. Impact, Monitoring and Management of Environmental Pollution. *Nova Science Publishers, Inc., Hauppauge, New York*. p. 638.
- EL NEMR, A. 2013. New Developments in Blue Biotechnology and Environmental Pollution Assessment. Nova Science Publishers Inc., Hauppauge, New York (Hard cover ISBN: 978-1-62808-138-1, e-book ISBN: 978-1-62808-139-8).
- EPA (ENVIEONMENTAL PROTECTION AGENCY). 2012. DDT A brief history and Status. http://www.epa.gov/pesticides/factsheets/chemicals/ddt-briefhistory-status.htm. Cited: 15 February 2014
- EQANI, S.A.M.A.S., MALIK, R.N., CINCINELLI, A., ZHANG, G., MOHAMMAD, A. and QADIR. 2013. Uptake of organochlorine pesticides (OCPs) and polychlorinated biphenyls (PCBs) by river water fish: the case of River Chenab. *Sci. Total Environ*. **450:** 83–91.
- ERIBE, E., ROSSELAND, B.O., BORGSTROM, R., SALBU, B., GEBREMARIAM, Z., DEDEBO, E., NORIL, H.R. and EKLO, O.M. 2011. Bioaccumulation of persistent organic pollutants (POPs) in fish species from Lake Koka, Ethoipia: The influence of lipid content and trophic position. *Sci. Total Environ.* **410-411**: 136-145.
- ESCO. 2005. Country situation report on persistent organic pollutants in Bangladesh. Dhaka, Bangladesh. Environment and Social Development Organization.
- FAO. 1999. FOA warns about the dangerous legacy of obsolete pesticides. Press release 99/31.
- FAO/WHO. 2010. Pesticide residues in food and feed. Acceptable daily intake; Codex Alimentarius Commission. FAO/WHO Food Standards, Rome.
- FAO/WHO. 2012. Codex Alimentarius Commission. Pesticide Residues in Food and Feed. Codex Pesticide Residues in Food Online Database.
- FATIMA, M. and KHAN, A.A. 1993. Cyclic changes in the gonads of *Rhingomugil corsula* (Ham) from river Yamuna India. *Asian fish. Sci*. **6**: 23-29.
- FERNANDEZ, P. and GRIMALT, JO. 2003. On the global distribution of persistent organic pollutants. *Chimia.* **57**: 514–21.
- FISK, A.T., HOBSON, K.A. and NORSTROM, R.J. 2001. Influence of chemical and biological factors on trophic transfer of persistent organic pollutants in the northwater polynya marine food web. *Environ. Sci. Technol*. **35**: 732–738.
- FISHBASE. A global information system on fishes. http://www.fishbase.org/, 2014.
- FRANTZEN, S., MAGE, A., IVERSEN, S. A. and JUSHAMN, K. 2011. Seasonal variation in the levels of organohalogen compounds in herring (*Clupea harengus*) from the Norwegian sea. *Chemosphere.* **85***:* 179-187.
- GASTON, C.P. 1986. Pesticide usage, registration and regulatory practices among selected countriesin Asia.
- GERBER, R., SMIT, N. J., VUREN, J.H.J.V., NAKAYAMAC, S.M.M., YOHANNES, Y.B., IKENAKA, Y., ISHIZUKA, M. and WEPENER, V. 2016. Bioaccumulation and human health risk assessment of DDT and other organochlorine pesticides in an apex aquatic predator from a premier conservation area. *Sci. Total Environ.* **550:** 522-533.
- GRUGDYEV, G.S., ZINCHEKO, V.A., KALININ, V.A. and SLOVTSOV, R.I. 1983. The Chemical Protection of Plants.UNEP (United Nations Environmental Protection), 2001. In: Final Act of the Conference of Plenipotentiaries on the Stockholm Convention on Persistent Organic Pollutants, Stockholm, Sweden.
- GUO, L., QIU, Y., ZHANG, G., ZHENG, G.J., LAM, P.K. and Li, X., 2008 (a). Levels and bioaccumulation of organochlorine pesticides (OCPs) and polybrominated diphenyl ethers (PBDEs) in fishes from the Pearl River estuary and Daya Bay, South China. *Environ. Pollut.* **152**: 604–611.
- GUO, Y, MENG, X.Z., TANG, H.L. and ZENG, E.Y. 2008 (b). Tissue distribution of organochlorine pesticides in fish collected from the Pearl River delta, China: Implications for fishery input source and bioaccumulation. *Environ. Pollut.* **155**: 150-156.
- GUPTA, S. and BANERJEE, S. 2016. *Eutropiichthys vacha* (Hamilton, 1822), a threatened fish of Indian subcontinent. *J. Fisheries.* **4(2)**: 397-400.
- GURUGE, K.and TANABE, S. 2001. Contamination by persistent organochlorines and butyltin compounds in the west coast of Sri Lanka. *Mar. Pollut. Bull.* **42**: 179–186.
- HASAN, M.N., ISLAM, H.M.R, AHMED, K.K.U., Y. MAHMUD,Y.and SIDDIQUEE, S. 2013. Screening and quantification of dichlorodiphenyltrichloroethane (DDT) and Dichlorovos in selected dry fish species of Bangladesh by GC-ECD detector. *Int. J. sci. res. manag.* **1(7)**: 352- 353.
- HASAN, M.N., ISLAM, H.M. R., AKTER, R., MAHMUD, Y., AHMED, K.K.U. and SIDDIQUEE, S. 2014. Determination of Dichlorodiphenyltrichloroethane Residues Levels in Commercial Marine Dry Fish from Different Regions of Bangladesh. *Annual rep. rev. biology* **4(17)**: 2722-2729.
- HAO, Q., SUN, Y.Z., XU, X.R., YAO, Z.W., WANGC, Y.S., ZHANG, Z.W., LUO, X.J. and MAI, B.X. 2014. Occurrence of persistent organic pollutants in marine fish from the Natuna Island, South China Sea. *Marine Pollut. Bull.* **85**: 274– 279.
- HEIDEN, T.K., HUTZ, R.J.and CARVAN, M.J., 2005. Accumulation, tissue distribution, and maternal transfer of dietary 2,3,7,8,-tetrachlorodibenzo-pdioxin: impacts on reproductive success of zebrafish. *Toxicol. Sci.* **87**: 497–507.
- HIES (Household Income and Expenditure Survey), for consumption: Report of the Household Income and Expenditure Survey 2010 (report published December 2011), Bangladesh Bureau of Statistics
- HOEKSTRA, P.F., OHARA, T.M. and FISK, A.T. 2003. Trophic transfer of persistent organochlorine contaminants (OCs) within an Arctic marine food web from the southern Beaufort-Chukchi Seas. *Environ. Pollut*. **124**: 509-522.
- HOLDEN, A.V. 1966. Organochlorine insecticide residues in salmonoid fish. *J. Appl. Ecol*. **3**: 45-53.
- HONG, S.H., YIM, U.H., SHIM, W.J., Oh, J.R. and LEE, I.S. 2003. Horizontal and vertical distribution of PCBs and chlorinated pesticides in sediments from Masan Bay, Korea. *Mar. Pollut. Bull*. **46:** 244–253.
- HOSSAIN, M. 1988. Nature and Impact of the Green Revolution in Bangladesh. Research Report No. 67. International Food Policy Research Institute, Washington, DC.
- HOSSAIN, M.A., SHOEB, M. and NAHAR, N. 2016. DDT and Its Metabolites in Fresh Water Fish Samples. *J. Food Sci. Engin*. **6**: 344-350.
- HOSSAIN, M.S., ROY, A, and RAHMAN, M. L. 2016. Food and feeding habit of Bele *Glossogobius giuris* (Hamilton and Buchannan, 1822) Collected from Mithamain Haor of Kishoreganj districts, northeastern Bangladesh. *Inter. J. Fish. Aquatic Stud.* **4(5)**: 84-88.
- HUED, A.C., LO NOSTRO, F.L.,WUNDERLIN, D.A. and BISTONI, M.A. 2013. Reproductive impairment of a viviparous fish species inhabiting a freshwater system with anthropogenic impact. *Arch. Environ. Contam. Toxicol*. **64 (2):** 281–290.
- IARC. 1974 . Some Organochlorine Pesticides. *IARC Monogr Eval Cercinog Risk Hum.* 5.
- IARC 1991. Occupational Exposurein Insecticide Application, and Some Pesticides. *IARC Monogr Eval Cercinog Risk Hum. 53.*
- ICPS 1976. DATA SHEETS ON PESTICIDES No. 21, DDT. World Health Organization (WHO).1979. Environmental Health Criteria 9, DDT and its derivatives. World Health Organization, Geneva.
- IPCS and IARC 2009. Human Health and Environmental Risk Evaluation of Selected Pops. http//www. (2009). kvk.pravara.com/ grapes MRL.pdf.
- IRAM, S., AHMAD, I., AHAD, K., MUHAMMAD, A. and ANJUM, S. 2009. Analysis of pesticide residues of Rawal and Simly Lakes. *Pak. J. Bot.* **41**: 1981– 1987.
- JAN, M.R., SHAH, J., KHAWAJA, M.A. and GUL, K. 2009. DDT residue in soil and water in and around the abandoned DDT manufacturing factory. *Environ. Monit. Assess*. **155:** 31–8.
- JARVINEN, A.W. and ANKLEY, G.T. 1999. Linkage of effects to tissue residues: development of a comprehensive database for aquatic organisms exposed to inorganic and organic chemicals. Society of Environmental Toxicology & Chemistry, Pensacola, FL. *For consumption: Report of the Household Income and Expenditure Survey 2010 (report published December 2011), Bangladesh Bureau of Statistics ii.*
- JAVEDANKHERAD, I., ESMILI-SARI, A. and BAHRAMIFAR, N. 2013. Levels and Distribution of Organochlorine Pesticides and Polychlorinated Biphenyls in Water and Sediment from the International Anzali Wetland, North of Iran. *Bull. Environ. Contam. Toxicol*. **90**: 285–290
- JAYASHREE, R. and VASUDEVN, N. 2007. Effect of tween 80 addedto the soil on the degradation of endosulfan by Pseudomonas aeruginosa. *Int. J. Environ. Sci. Tech*. **4 (2)**: 203-210.
- JHINGRAN, A.G.1961. Studies on the maturity and fecundity of the Gangatic unchovy, *Setipinna phase* (Ham). *Indian J. Fish.* **8**: 291-311.
- JIANG, Y.F., WANG, X.T., JIA. Y., WANG, F., WU, M.H., SHENG, G.Y. and FU, J.M. 2009. Occurrence, distribution and possible sources of organochlorine pesticides in agricultural soil of Shanghai, China. *J. Hazard Mater.* **170**: 989– 997.
- KANNAN, K., TANABE, S., TATSUKAWA, R. and SINHA, R. 1994. Biodegradation capacity and residue pattern of organochlorines in Ganges river dolphins from India. *Toxicol. Environ. Chem.* **42:** 249–261.
- KARLSSON, H., MUIR, D.C.G., TEIXIERA, C.F., BURNISTON, D.A., STRACHAN, W.M.J., HECKY, R.E., MWITA, J., BOOTSMA, H.A., GRIFT, N.P., KIDD, K.A. and ROSENBERG, B. 2000. Persistent chlorinated pesticides in air, water and precipitation from the Lake Malawi area, Southern Africa. *Environ. Sci. Technol*. **34**: 4490–4495.
- KEITH, E.H. 1997. Environmental Endocrine Disrupter. *New York: John Willy & Sons.*
- KELLY, B.C., IKONOMOU, M.G., BLAIR, J.D. and GOBA, F.A.P.C. 2008. Hydroxylated and methoxylated polybrominated diphenyle ethers (PBDES) in a Canadian Arctic marine foodweb. *Environ. Sci. Technol*. **42**: 7069-72.
- KENNISH, M.J. 1997. Practical Handbook of Estuarine and Marine Pollution. Halogenated Hydrocarbons. CRC Press, Marine sciences series Boca Raton.
- KENT, J.C. and JOHNSON, D.W. 1979. Organochlorine residues in fish, water, and sediment of American Falls Reservoir, Idaho. *Pestic. Monit. J.* **13(1)**: 28–34.
- KHAN, M.A. 1988. Biology of *Labeo calbasu* (Ham-Buch) from Trilaiya reservoiv, Bihar. Length weight relationship, condition index and feeding habits. *Proc. Nat. Acad. Sci. India.* **58**: 41-47.
- KHAN, M.S., AMBAK, M.A. and MOHSIN, A.K.M. 1988. Food and feeding biology of atropical catfish *Mystus nemurus*, with reference to its functional morphology. *Indian J. Fish.***35***:* 78-84.
- KIDD, K.A., BOOTSMA, H.A., HESSLIEIN, R.H. and MUIR, D.C.G. 2000. Biomagnification of DDT through the benthic and pelagic food webs of Lake

Malawi, East Africa, importance of trophic level and carbon source. *Environ. Sci. Technol.* **35**: 14-20.

- KIDD, K.A., SCHINDLER, D.W., HESSLEIN, R.H. and MUIR, D.C.G. 1998. Effects of trophic position and lipid on organochlorine concentrations in fishes from subarctic lake in the Canadian Arctic. *Environ. Pollut.* **102:** 91-103.
- KIDWELL, J. M., PHILLIPS, L. J. and BIRCHARD, G. F. 1990. Comparative analysis of contaminants level in bottom feeding and predatory fish using the national contaminants biomonitoring program data. *B. Environ. Contam. Tox.* **55(6)**: 919-923.
- KIRAN, V.U.G. and WAGHRAY, S. 1998. Food and feeding habits of Notopterous notopterous (Pallas) of Saroornagar lake, Hyderabad. *Indian J. Fish.* 45(3): 355-359.
- KONG, K.Y., CHEUNG, K.C., WONG, C.K.C and WONG, M.H. 2005. The residual dynamic of polycyclic aromatic hydrocarbons and organochlorine pesticides in fishponds of the pearl River delta, South China. *Water Res.* **39**: 1831-43.
- KONWICK, B.J., GARRISON, A.W., BLACK, M.C., AVANTS, J.K. and FISK, A.T. 2006. Bioaccumulation, biotransformation, and metabolite formation of fipronil and chiral legacy pesticides in rainbow trout. *Environ. Sci. Technol*. **40**: 2930–2936.
- KRIEGER, N. W. M., HIATT, R.A., RIVERA, M., VOGELMAN, J., ORENTREICH, N. 1994. Breast cancer and serum organochlorines: a prospective study among white, black and Asian women. *J. Natl. Cancer Inst*. **86**: 589-599.
- KUMARI, B., MADAN, V. and KATHPAL, T. 2007. Pesticide residues in rain water from Hisar, India. *Environ. Monit. Assess*. **133**: 467–471.
- LAW, R.J., ALAEE, M., ALLCHIN, C.R., BOON, J.P., LEBEUF, M. and LEPOM, P. 2003. Levels and trends of polybrominated diphenylethers and other brominated flame retardants in wildlife. *Environ. Int*. **29**: 757–770.
- LEE, D.H., LEE, I. K., Song, K., STEFFES, M., TOSCANO, W., BAKER, B.A. and JACOBS, D.R. 2006. A Strong Dose-Response Relation Between Serum Concentrations of Persistent Organic Pollutants and Diabetes. *Diabetes Care* **29**: 1638-1644.
- LI, X., GAN, Y., YANG, X., ZHOU, J., DAI, J. and XU, M. 2008. Human health risk of organochlorine pesticides (OCPs) and polychlorinated biphenyls (PCBs) in edible fish from Huairou Reservoir and Gaobeidian Lake in Beijing, China. *Food. Chem.* **109**: 348–354.
- LOHMANN, R., BREIVIK, K., DACHS, J. and MUIR, D. 2007. Global fate of POPs: Current and future research directions. *Environ. Pollut*. **150**: 150–65.
- LONGMECKER, P. M., KLEBANOFF, A. M., DUNSON, B. D., GUO, X., CHEN, Z., ZHOU, H. and BROCK, W. J. 2005. Maternal serum level of the DDT metabolite DDE in relation to fetal loss in previous pregnancies. *Environ. Res.* **97**: 127 – 133.
- LONGNECKER, M. P. R., W. J.and LUCIER, G. 1997. The human health effect of DDT(dichlorodiphenylethane) and PCBs (polychlorinated biphenyls) and an overview og organochlorines in public health. *Annu. Rev. Public Health* **18:** 211-244.
- LONGNECKER, M. P., KLEBANOFF, M. A., ZHOU, H. and BROCK, J. W. 2001. ssociation between maternal serum concentration of the DDT metabolite DDE and preterm and small-for-gestational-age babies at birth. *Lancet.* **358 (9276)**: 110-114.
- MA, M., FENG, Z., GUAN, C., MA, Y., XU, H. and LI, H. 2001. DDT, PAH and PCB in sediments from the interdial zone of the Bohai Sea and the Yellow Sea. *Mar. Pollut. Bull*. **42**: 132–136.
- MAHMUD, M.N.U., KHALIL, F., RAHMAN, M.M., MAMUN, M.I.R., SHOEB, M., EL-ATY, A.M., PARK, J.H., SHIN, H.C., NAHAR, N. and SHIM, J.H. 2015. Analysis of DDT and its metabolites in soil and water samples obtained in the vicinity of a closed-down factory in Bangladesh using various extraction methods. *Environ. Monit. Assess*. **187**: 743 (1-12).
- MAMUN, M. I. R., ZAMIR, R., NAHAR, N., MOSIHUZZAMAN, M., LINDERHOLM, L., ATHANASIADOU, M. and BERGMAN, A. 2007. Traditional Organochlorine pollutants in blood from humans living in Bangladesh capital area. *Organohalogen Comp.* **69:** 2026-2030.
- MANIRAKIZAA, P., COVACIA, A., NIZIGIYMANAB, L., NTAKIMAZIB, G. and SCHEPENS, P. 2002. Persistent chlorinated pesticides and polychlorinated

biphenyls in selected fish species from Lake Tanganyika, Burundi, Africa. *Environ. Pollut.* **117**: 447–455

- MANOJKUMAR, P.P. and ACHARYA, P. 1990. Morphometriy, Length-weighi relationship and food and feeding habits of *Otolithoides biauritus* (Cantor, 1805) of Bombay waters. *J. Ind. Fisher. Assoc.* **20:** 31-36.
- MANSOUR, S.A. 2009. Persistent Organic Pollutants (POPs) in Africa: Egyptian scenario. *Hum Exp Toxicol*. **28**: 531–66.
- MASTOVSKA, K., DORWEILER, K. J., LEHOTAY, S. J., WEGSCHEID, J. S., and SZPYLKA, K. A. 2010. Pesticide Multiresidue Analysis in Cereal Grains Using Modified QuEChERS Method Combined with Automated Direct Sample Introduction GC-TOFMS and UPLC-MS/MS Techniques. *J. Agric. Food Chem*. **58**: 5959–5972.
- MATIN, M., MALEK, M., AMIN, M., RAHMAN, S., KHATOON, J. and RAHMAN, M. 1998. Organochlorine insecticide residues in surface and undergroundwater from different regions of Bangladesh. *Agri. Ecosys. Environ*. **69**: 11–15.
- MINTEN, B., ALAM, A. Z. M. S., DEB, U. K., KABIR, A. Z. K., LABORDE, D., HASSANULLA, M. and MURSHID, K. A. S. 2010. Agricultural Marketing, Price Stabilization, Value Chains, and Global/Regional Trade. *Bang. Food Secur. Invest. Forum Dhaka.*
- MISCHKE, T., BRUNETTI, K., ACOSTA, V., WEAVER, D. and Brown, M. 1985. Agriculture source of DDT residues in California's environment. California, US: California Department of Food and Agriculture.
- MISHRA, K., SHARMA, R.C. and KUMAR, S. 2012. Contamination levels and spatial distribution of organochlorine pesticides in soils from India. *Ecotoxicol. Environ. Saf*. **76**: 215–225.
- MONIRITH, I., NAKATA, H., TANABE, S. and SEANG, T.T. 1999. Persistent organochlorine residues in marine and freshwater fish in Cambodia. *Mar. Pollut. Bull*. **38**: 604–612.
- MORGAN, E.J. and LOHMANN, R. 2010. Dietary uptake from historically contaminated sediments as a source of PCBs to migratory fish and invertebrates in an urban estuary. *Environ. Sci. Technol.* **44**: 5444–5449.
- MUIR, D., Riget, F., *et al*. 2000. Circumpolar Trends of PCBs and Organochlorine Pesticides in the Arctic Marine Environment Inferred from Levels in Ringed Seals. *Environ. Sci.Technol.* **34**: 2431.
- MUIR, D.C.G., FORD, C.A., GRIFT, N.P.,METNER, D.A. and LOCKHART, W.L. 1990. Geographical variation of chlorinated hydrocarbons of Burbot (*Lota lota*) from remote and rivers in Canada. *Arch. Environ. Contam. Taxicol.* **19**: 530-42
- MURALIDHARAN, S., DHANANJAYAN, V. and JAHANTHI, P. 2009. Organochlorine pesticides in commercial marine fishes of Coimbatore, India and their suitability for human consumption. *Environ. Res*. **109**: 15–21.
- MUSTAFA, G., AHMED, A.T.A. and ISLAM, K.R. 1980. Food, feeding habits and fecundity of freshwater perch, Meni fish. *Bangldesh J Agri*. [http://agris.fao.org/arris-](http://agris.fao.org/arris)search /search. do ?record ID =XB812540
- MYERS, M.J., YLTALO, G.M., KRAHN, M.M., BOYD, D., CALKINS, D. and BURKANOV, V. 2008. Organochlorine contaminants in endangered steller sea lion pups (*Eumetopias jubatus*) from western Alaska and the Russian Far East. *Sci. Total Environ*. **396**: 60-69.
- NADAL, M., SCHUHMACHER, M and DOMINGO, J.L. 2004. Levels of PAHs in Soil and Vegetation Samples from Tarragora County, Spain. *Environ. Poll.* **132**: 1-11.
- NAHAR, N., MAMUN, M.I.R., ZAMIR, R. and. MOSHIUZZAMAN, M. 2008. Analysis of Pesticide Residues in Some Local Fish and Vegetable. *Dhaka Univ. J. Sci.* **56 (2)**: 1-4.
- NAKATA, H., KAWAZONO, M., ARIZONO, K., ABE, S., KITANO, T.and SHINADA, H. 2002. Organochlorine pesticides and polychlorinated biphenyl residues in foodstuffs and human tissues from China: status of contamination, historical trend and human dietary exposure. *Arch. Environ. Contam. Toxicol*. **43**: 473-480.
- NDS (Narayangang District Statistics). 2013. Reported by the Bangladesh Bureau of Statistics (BBS), Statistics and Informatics Division (SID), Ministry of Planning, Government for the People's Republic of Bangladesh.
- NFW (National Fisheries Week). 2015. Compendium (in Bangla) Department of Fisheries, Ministry of Fisheries and Livestock, Bangladesh. pp:13,133-134.
- NISHINA, T., KIEN, C.N., NOI, N.V., NGOC, H.M., KIM, C.S., TANAKA, S. and IWASAKI, K. 2010. Pesticide residues in soils, sediments, and vegetables in the Red River Delta, northern Vietnam. *Environ. Moni.t Assess*. **169**: 285–297.
- NNAMUYOMBA, P., MBABAZI, J. and NTALE, M. 2014. Dichloro-diphenyletrichloroethane (DDT) residue levels in marketed Silver fish (*Rastreneobole argentea*) caught from major water bodies in Uganda. *African J. Pure Appl. Chem*. **8**(6): 94-101.
- NORENA-BARROSO, E., SIMA-ALVAREZ, R., GOLDd-BOUCHOT, G. and ZAPATA-PEREZ, O. 2004. Persistent organic pollutants and histological lesions in Mayan catfish Ariopsis assimilis from the bay of Chetumal, Mexico. *Mar. Pollut. Bull*. **48:** 263–269.
- NORMAN D. SCHOENTHAL 1963. Some Effects of DDT on cold -water fish and fish-food organisms. A thesis DOCTOR OF PHILOSOPHY in Fish and Wildlife Management MONTANA STATE COLLEGE Bozeman, Montana.
- NORSTROM, R. J., SIMON, M., MUIR, D.C.G. and Schweinsburg, R.E. 1998. Organochlorine contaminants in Arctic marine food chains: Identification, geographical distribution and temporal trends in polar bears. *Environ.Sci. Technol.* **22**: 1063-1071.
- NTOW, J.W. 2005. Pesticide Residues in Volta Lake, Ghana. *Lakes Reservoirs Res. Manag.* **10**: 243-248.
- NTOW, W.J. 2001. Organochlorine pesticide in water, sediments, crops and human fluids in a farming community in Ghana. Arch. *Environ. Contam. Toxicol.* **40:** 557-563.
- NUAPIA, Y., CHIMUKA, L. and CUKROWKA, E. 2016. Assessment of organochlorine pesticide residues in raw food samples from open markets in two African cities. *Chemosphere* **164:** 480-487
- NYHOLM, J.R., NORMAN, A., NORRGEN, L., HAGLUND, P., ANDERSSON, P.L. 2008. Maternal transfer of brominated flame retardants in zebrafish (Danio rerio). *Chemosphere* **73**: 203–208.
- ONDARZA, P.M., GONZALEZ, M., FILLMANN, G. and MIGLIORANZA, K.S.B. 2011. Polybrominated diphenyl ethers and organochlorine compound levels in

brown trout (Salmo trutta) from Andean Patagonia, Argentina. *Chemosphere* **83**: 1597–602.

- ONDARZA, P.M., GONZALEZ, M., FILLMANN, G. and MIGLIORANZA, K.S.B. 2012. Increasing levels of persistent organic pollutants in rainbow trout (*Oncorhynchus mykiss*) following a mega-flooding episode in the Negro River basin, Argentinean Patagonia. *Sci. Tot. Environ*. **419**: 233–239.
- ONDARZA, P.M., GONZALEZ, M., FILLMANN, G.and MIGLIORANZA, K.S.B. 2011. Polybrominateddiphenyl ethers and organochlorine compound levels in brown trout Salmo trutta from Andean Patagonia. *Chemosphere* **83**: 1597–1602.
- OSAFO, A.S. and FREMPONG, E. 1998. Lindane and endosulfan residues in water and fish in the Ashanti region of Ghana. *J.Ghana Sci. Assoc.* **1 (1)**: 135-140.
- PASTOR, D., BOIX, J. and FERNANDEZ, V. 1996. Bioaccumulation of organochlorined contaminants in three estuarine fish species (*Mullus Barbatus, Mugil Cephalus* and *Dicentrarcus Labrax*). *Mar. Pollut. Bull*. **32:** 257-262.
- PARVEEN, S. and NAKAGOSHI, N. 2001. An analysis of pesticide use for rice pest management in Bangladesh. *J. Int. Develop. Cooper.* **8**: 107–126.
- PAZOU, E.Y.A., LALEYE, P., BOKO, M., VAN GESTEL, C.A.M., AHISSOU, H., AKPONA, S., VAN HATTUM, B., SWART, K. and VAN STRAALEN, N.M. 2006. Contamination of fish by organochlorine pesticide levels in the Ouémé River catchment in the Republic of Bénin. *Environ. Inter.* **32:** 594–599.
- PETERS, A.J., JONES, K.C., FLOWER, R.J., APPLEBY, P.G., RAMDANI, M., KRAIEM, M.M. 2001. Recent environmental change in North African wetland lakes: A baseline study of organochlorine contaminant residues in sediments from nine sites in the CASSARINA project. *Aquat. Ecol*. **35**: 449–59.
- PHILLIPS, D. J. H. 1980. Quantitative Aquatic Biological Indicators. *Applied Science Publishers, London.* pp. 185-187.
- POMA, G., VOLTA, P., ROSCIOLI, C., BETTINETTI, R. and GUZZELLAOM, L. 2014. Concentrations and trophic interactions of novel brominated flame retardants, HBCD, and PBDEs in zooplankton and fish from Lake Maggiore (Northern Italy). *Sci. Total Environ.* **481**: 401–408.
- PRETTY, J. and HINE, R. 2005. Pesticide Use and the Rnvironment. London: Earthscan. Pesticide Action Network (2012) The DDT story.

http://www.panna.org/issues/persistent-poisons/the-ddt-story. Cited: 27 February 2014.

- RAHMAN, A. K. A. 2005. *Freshwater fishes of Bangladesh*. 2nd ed. Zoological Society of Bangladesh, Dhaka. pp-79, 141, 183, 305, 270-273, 203, 333, 130,133, 176,285, 286, 345, 346, 354, 355, 338.
- RAHMAN, M. M. 2000. Pesticides: Their use and Problems in Context of Bangladesh. Workshop on Conventional and Nuclear Technique for Pesticide Residue Studies in Food and Environment, IFRB,Savar, Bangladesh.
- RAHMAN, M., MALEK, M. and MATIN, M. 1995. Trend of pesticide usage in Bangladesh. *Sci. Total. Environ.* **159**: 33–9.
- RAHMAN, M., MALEK, M. and MATIN, M. 1995. Trend of pesticide usage in Bangladesh. *Sci. Total Environ.* **159**: 33–39.
- RAHMAN, M.S., RAHMAN, M.M., PARVEZ1, M.S. and RASHED-UN-NABI, M. 2016. Feeding habit and length-weight relationship of a Mudskipper *Apocryptes bato* (Hamilton, 1822) from the coast of Chittagong, Bangladesh. *J. Bang. Acad. Sci.* **40(1)**: 57-64.
- RAHMAN, S. 2005. Environmental impacts of technological change in Bangladesh agriculture: farmers' perceptions, determinants, and effects on resource allocation decisions. *Agri. Eco.* **33**: 107–116.
- RAHMAN, S. and THAPA, G.B. 1999. Environmental impacts of technological change in Bangladesh agriculture: farmers' perceptions and empirical evidence. *Outlook Agri.* **28**: 233–238.
- RAJENDRAN, R.B., IMAGAWA, T., TAO, H. and RAMESH, R. 2005. Distribution of PCBs, HCHs and DDTs, and their ecotoxicological implications in Bay of Bengal, India. *Environ. Int.***31**: 503–12.
- RAMESH, A., TANABE, S., KANNAN, K., SUBRANANIAN, A., KUMARAN, P. and TATSUKAWA, R. 1992. Characteristic trend of persistent organochlorine contamination in wildlife from a tropical agricultural watershed, South India. *Arch. Environ. Contam. Toxicol.* **23**: 26–36.
- RAO, T., VIVEK, C., REDDY, A., VEERAIAH, K. and PADMVATHI, P. 2014. Determination of organochlorine pesticide residues in the water of paddy fields

of Prakasam district, Andhra Pradesh, India. *Int. J. Rec. Sci. Res.* **5(11)**: 1983- 1987.

- RASUL, G. and G. THAPA. 2003. Sustainability Analysis of Ecological and Conventional Agricultural Systems in Bangladesh. *World Develop.* **31(10)**: 1721-1741.
- REINERT, R. E., STONE, L. J., and WILLFORD, W. A. 1974. Effects of temperature on residues in fodder from rural areas of Haryana, India. *Toxicol. Environ. Chem*. **95:** 69–81.
- [RISSATO,](https://www.researchgate.net/profile/Sandra_Rissato) S., [MARIO, S., GALHIANE,](https://www.researchgate.net/researcher/76351082_Mario_S_Galhiane) [VALDECIR, F., XIMENES,](https://www.researchgate.net/profile/Valdecir_Ximenes) [A. and](https://www.researchgate.net/researcher/28326659_Aline_A_Cavalari) [CAVALARI,](https://www.researchgate.net/researcher/28326659_Aline_A_Cavalari) A. 2007. Organochlorine pesticides and polychlorinated biphenyls in soil and water samples in the Northeastern part of São Paulo State, Brazil. *Chemosphere* **65(11)**:1949-1958.
- RITTER, L., SOLOMON, R. K. and FORGET, J. 1995. Assessment Report on: DDT-Aldrin-Dieldrin-Chlordane-Heptachlor-Haxachlorobenzene-Mirex-Toxaphene-Polychlorinated Biphenyls-Dioxin and Furans. For: *The International program on Chemical Safty (IPCS)* within the framework of the Inter-organization program for the Sound Management of Chanicals (IOMC)**:** 24.
- ROBINSON, T., ALI, U., MAHMOOD, A., CHAUDHRY, M.J.I., LI, J., ZHANG, G., JONES, K.C. and MALIK, R.N. 2016 Concentrations and patterns of organochlorines (OCs) in various fish species from the Indus River, Pakistan: A human health risk assessment. *Sci. Total Enviorn*. **541**: 1232-1242.
- ROOS, N., WAHAB, M.A., HOSSAIN, M.A.R. and THILSTED, S.H. 2007. Linking human nutrition and fisheries: incorporating micronutrient dense, small indigenous fish species in carp polyculture production in Bangladesh. *Food Nutrit. Bull.* **28 (2):** 280–293.
- RUSSELL, R.W., GOBAS and F., HAFFNER, G.D. 1999. Maternal transfer and in ovo exposure of organochlorines in oviparous organisms: a model and field verification. *Environ. Sci. Technol*. **33**: 416–420.
- SABLJIC, A. 2001. QSAR models for estimating properties of persistent organic pollutants required in evaluation of their environmental fate and risk. *Chemosphere* **43:** 363–375.
- SAID, K., TAREK, O.M., El MOSELHY, ABDELAZIW, M., RASHAD, SHREASAH, M. A. 2008. Organochlorine Contaminants in Water, Sediment and Fish of Lake Burullus, Egyptian Mediterranean Sea. *[Bulletin of Environ.](http://link.springer.com/journal/128) [Contam.](http://link.springer.com/journal/128) Toxicol.* **81(2)**: 136–146.
- SCHMID, P., KOHLER, M., GUJER, E., ZENNEGG, M. and LANFRANCHI, M. 2007. Persistent organic pollutants, brominated flame retardants and synthetic musks in fish from remote alpine lakes in Switzerland. *Chemosphere* **67:** S16– S21.
- SCHNITZLER, J.G., THOM E, J.M., LEPAGE, M. and DAS, K. 2011.Organochlorine Pesticides, Polychlorinated Biphenyls and Trace Elements in Wild European Sea Bass (Dicentrarchus labrax) off European Estuaries. *Sci. Total. Environ.* **409**: 3680-3686.
- SENTHILKUMAR, K., KANNAN, K., SUBRAMAIAN, A. and TANABE, S. 2001. Accumulation of organochlorine pesticides and polychlorinated biphenyls in sediments, aquatic organisms, birds, bird eggs and bat collected from south India. *Environ. Sci. Pollut. Res.***8:** 35–47.
- SERAJUDDIN, M.A. KHAN, A. and MUSTAFA, S. 1998. Food and feeding habits of the spiny eel. *Mastacembelus armatus. Asian fish. Sci*. **11**: 271-278.
- SETHI, P., BHATTACHARYYA, A. and SRAKAR, A. 1999. Current trends of some organochlorinated pesticides in Yamuna River sediments around Delhi. *Environ. Pollut. Cont.* **32:** 234-245.
- SHAFI, M. and QUDDUS, M.M.A. 1982. *Bangladesher Matshya Shampad (in* Bengali). 1st ed. Bangla Academy, Dhaka. pp-2-14.
- SHARMA, H.R., KAUSHIK, A. and KAUSHIK, C.P. 2013. Organochlorine pesticide residues in fodder from rural areas of Haryana, India. *Toxico.l Environ. Chem*. **95**: 69–81.
- SHOU, R., ZHU, I. and KONG, Q. 2007. Persitant chlorinated pesticides in fish species from Qiantang River in East China. *Chemosphere* **68**: 838-47.
- SINGH, P.B. and SING, V. 2008. Pesticide bioaccumulation and plasma sex steroids in fishes during breeding phase from North India. *Environ. Toxicol. Pharmacol.* **25:** 342-350.
- SINGH, S., BHUTIA, D., SARKAR, S., RAI, B. K., PAL, J., BHATTACHARJEE, S. and BAHADUR, M. 2015. Analyses of pesticide residues in water, sediment and fish tissue from river Deomoni flowing through the teagardens of Terai Region of West Bengal, India. *Internat. .J. Fisher. Aqua. Std*. **3(2)**: 17-23
- SMITH, A.G. 1991. Chlorinated hydrocarbone insecticides. In Handbook of Pesticides Toxicology (Haywas WJ, Laws ER, Eds). *San Deigo/ New York: Academy Press Inc.* pp- 731-915.
- SSEBUGERE, P., KIREMIRE, B.T., KISHIMBA, M., WANDIGA, S.O., NYANZI, S.A. and WASSWA, J. 2009. DDT and metabolites in fish from Lake Edward, Uganda. *Chemosphere*. **76:** 212-215.
- SUN, R.X., LUO, X.J., TAN, X.X., TANG, B., Li, Z.R. and MAI, B.X. 2015. Legacy and emerging halogenated organic pollutants in marine organisms from the Pearl River Estuary, South China. *Chemosphere* **139:** 565–571.
- SUN, Y.X., HAO, Q., XU, X.R., LUO, X.J., WANG, S.L., ZHANG and Z.W., MAI, B.X. 2014. Persistent organic pollutants in marine fish from Yongxing Island, South China Sea: Levels, composition profiles and human dietary exposure assessment. *Chemosphere* **98:** 84–90.
- SUSKIND, R. R. 1977. Environment and the skin. *Environ. Hlth. Persp.* **20**: 27-37.
- SVERDRUP, L.E., NIELSEN, T. and KROGH, P.H. 2002. Soil ecotoxicity of polycyclic aromatic hydrocarbons in relation to soil sorption, lipophilicity and water solubility. *Environ. Sci. Technol*. **36:** 2429-2435.
- TAKADA, N., KAWASHIMA, H. and OHGA, K. 2003. An inquiry into current status of DDT use in Bangladesh. *World Res. Rev.* **15(2):** 217-225.
- TAKAYAMA, S. S. S., DALGARD, D.W., THORGEIRSON, U.P. and ADAMSON, R.H. 1999. Effect of long -term oral administration of DDT on nonhuman primates. *J. Cancer Res. Clin. Oncol.* **125:** 219-225.
- TALUKDER, R.K. 2005. Food security, self-sufficiency and nutrition gap in Bangladesh. *Bangladesh Develop. Stud.* **31 (3&4):** 35–62.
- THUY, T.T. 2015. Effects of DDT on environment and human health. *J. Educat. Social Sci.* **2 (Oct.):**108-114.
- TOMATIS, E. H. J. 2000. Evidence of carcinogenicity of DDT in nonhumanprimates. *J. Cancer Res. Clin. Oncol*. **126:** 246-250.
- TOMATIS, L. T. V. 1975. Studies of carcinogenicity of DDT. *Gann Monogr* **17**: 219-241.
- TOMATIS, L. T. V., CHARLES, R.T., BOIOCCHI, M. and GATI, E. 1974. Effect of long-term exposure to 1,1-dichloro-2,2-bis (*p-*chlorophenyl)ethylene to 1,1 dichloro-2,2-bis (*p-*chlorophenyl)ehane, and to the two chemicals combined on CF-1 mice. *J. Natl. Cancer Inst.* **52:** 883-891.
- TURUSOV, V.S. D. N., TOMTIS, L., GATI, E. and CHARLES, R.T. 1973. Tumors in CF-1 mice exposed for six consecutive generations to DDT. *J Natl. Cancer Inst*. **51:** 983-997.
- ULRICH, B. 1983. A concept of forest ecosystem stability and of acid deposition as driving force for destabilization. In: Ulrich, B., and PankvaIls, J. (eds), *Effect of accumulation of air pollutants in forest ecosystem,* pp. 1-19. D. Riedel Publishing Co., Holland.
- UNEP. 2001. Final act of the conference of plenipotentiaries on the Stockholm convention on persistent organic pollutants. Stockholm, Sweden, United Nations Environment programme.
- UPADHI, F. and WOKOMA, O. A. F. 2012. Examination of some pesticide residues in surface water, sediment and fish tissue of Elachi creek, Niger Delta, Nigeria. *Res. J .Environ. Earth Sci.* **4(11):** 939-944.
- VALLACK, H.W., BAKKER, D.J., BRANDT, I., BROSTOM-LUNDEN, E., BROUWER, A. and BULL, K.R. 1998. Controlling persistent organic pollutants–What next? *Environ. Toxicol. Phamacol*. **6:** 143–175.
- VAN WENDEL, D.J.B., WESSEELING, C.K.H., MOGE, P., GARCIA, M. and MERGLER, D. 2001. Chronic nervous-system effect of long term occupational exposuer to DDT. *Lancet* **357:** 1014-1016.
- VERHAERT, V., COVACI, A., BOUILLON, S., ABRANTES, K., MUSIBONE, D., BERVOETS, L., VERHYENE, E., BLUSTA, E. 2013. Baseline levels and trophic transfer of persistent organic pollutants in sediments and biota from the Congo River Basin (DR Congo). *Environ. Intern.* **59:** 290-302.
- VERREAULT, J., LETCHER, R.J., MUIR, D.C.G., CHU, S.G., GEBBINK, W.A. and GABRIELSEN, G.W. 2005. New organochlorine contaminants and

metabolites in plasma and eggs of glyaucous gulls (*Larus hyperboreus*) from the Norwegian Arctic. *Environ. Toxicol. Chem*. **24**: 2486-2499.

- VITALE, F., *et al*. 201 2. Serum concentrations of persistent organic pollutants (POPs) in the inhabitants of a Sicilian city. *Chemosphere* **89**: 970–974.
- VOLLENWEIDER, J. J., HEINTZ, R. A., SCHAUFLER, L. and BRADSHAW, R. 2011. Seasonal cycles in whole-body proximate composition and energy content of forage fish vary with water depth. *Mar Biol*. **158:**413–427.
- VORKAMP, K., THOMSEN, M., MOLLER, S., FALK, K. and SORENSEN, P.B. 2009. Persistent organochlorine compounds in peregrine falcon (*Falco peregrinus*) eggs from South Greenland: levels and temporal changes between 1986-2003. *Environ. Internat.* **35**: 336-341.
- WALKER, K., VALLERO, D.A. and LEWIS, R.G. 1999. Factors influencing the distribution of lindane and other hexachlorocyclohexanes in the environment. *Environ. Sci. Technol.* **33**: 4373-4378.
- WANDIGA, M. O. S. O. AND JUMBA, I. O. 2002. The status of persistent organic pollutants in Lake Victoria catchment. **34:** 534-540.
- WANDIGA, S. 2001. Use and distribution of organochlorine pesticides. The future in Africa. IUPAC. *Pure Appl. Chem*. **73:** 1147-55.
- WANG, W., HUANG, M.J., WU, F.Y., KANG, Y., WANG, H.S., CHEUNG, K.C. and WONG, M.H. 2013. Risk assessment of bioaccessible organochlorine pesticides exposure via indoor and outdoor dust. *Atmos Environ.* 77: 525–533.
- WANIA, F. and MACKAY, D. 1996. Tracking the distribution of persistent organic pollutants. *Environ. Sci. Technol*. **30:** 390A–396A.
- WASSERMAN, M.T. L., WASSERMAN, D., DAY, N.E., GRONER, Y., LAZARIVICI, S., ROSENFELD, D. 1974. Epidemiology of organochlorine insecticides in the adupose tissue of Israells. *Pestic. Monit. J.* **8:** 1-7.
- [WEPENER,](file:///C:\Users\Azam\Desktop\New%20paper\Seasonal%20Bioaccumulation%20of%20Organohalogens%20in%20Tigerfish,%20Hydrocynus%20vittatus%20Castelnau,%20from%20Lake%20Pongolapoort,%20South%20Africa%20_%20SpringerLink.htm%23author-details-1) V., [SMIT,](file:///C:\Users\Azam\Desktop\New%20paper\Seasonal%20Bioaccumulation%20of%20Organohalogens%20in%20Tigerfish,%20Hydrocynus%20vittatus%20Castelnau,%20from%20Lake%20Pongolapoort,%20South%20Africa%20_%20SpringerLink.htm%23author-details-2) N., [COVACI,](file:///C:\Users\Azam\Desktop\New%20paper\Seasonal%20Bioaccumulation%20of%20Organohalogens%20in%20Tigerfish,%20Hydrocynus%20vittatus%20Castelnau,%20from%20Lake%20Pongolapoort,%20South%20Africa%20_%20SpringerLink.htm%23author-details-3) A., [DYKE, S.](file:///C:\Users\Azam\Desktop\New%20paper\Seasonal%20Bioaccumulation%20of%20Organohalogens%20in%20Tigerfish,%20Hydrocynus%20vittatus%20Castelnau,%20from%20Lake%20Pongolapoort,%20South%20Africa%20_%20SpringerLink.htm%23author-details-4) and [BEVOETS, L. 2](file:///C:\Users\Azam\Desktop\New%20paper\Seasonal%20Bioaccumulation%20of%20Organohalogens%20in%20Tigerfish,%20Hydrocynus%20vittatus%20Castelnau,%20from%20Lake%20Pongolapoort,%20South%20Africa%20_%20SpringerLink.htm%23author-details-5)012. Seasonal Bioaccumulation of Organohalogens in Tigerfish, *Hydrocynus vittatus* Castelnau, from Lake Pongolapoort, South Africa. *[Bull. Environ. Contam.](http://link.springer.com/journal/128) [Toxicol](http://link.springer.com/journal/128)*. **8[8\(2\)](http://link.springer.com/journal/128/88/2/page/1)**: 277–282.
- WHO (World Health Organization). 1973. Safe use of pesticide. WHO Technical Report. S. N. 513. Geneva: World Health Organization
- WHO (World Health Organisation). 2010. DDT in indoor residual spraying: Human heath aspects. Environmental health criteria 241.
- WHO (World Health Organization). 1979. DDT and its derivatives. Environmental Health Criteria 9. Geneva, Switzerland: World Health Organization.
- WOLF, M.S. C. G., BARRETT, J.C. and HUFF, J. 1996. Breast cancer and environmental riskfactors:epidemiological and experimental findings. *Ann. Rev. Pharmacol.Toxicol*. **36**: 573-596.
- WOLF, M.S. T. P., LEE, E.W., RIVERA, M. and DUBIN, N. 1993. Blood levels of organochlorine residues and risk of breast cancer. *J. Natl. Cancer Inst*. **85**: 648-652.
- XIA, C.H., LAM, J.C.W., WU, X.G., XIE, Z.Q. and LAM, P.K.S. 2012. Polychlorinated biphenyls (PCBs) in marine fishes from China: levels, distribution and risk assessment. *Chemosphere.* **89:** 944–949.
- YANG, R., WANG, Y., LI, A., ZHANG, Q., JING, C. and WANG, T. 2010. Organochlorine pesticides and PCBs in fish from lakes of the Tibetan Plateau and the implications. *Environ. Pollut.* **158:** 2310–2316.
- YIM, U., HONG, S., SHIM, W. and OH, J. 2005. Levels of persistent organochlorine contaminants in fish from Korea and their potential health risk. Arch. *Environ. Contam. Toxicol.* **48:** 358–366.
- YLITALO, G.M., MATKIN, C.O., BUZITIS, J., KRAHN, M.M., JONES, L.L. and ROWLES, T. 2001. Influence of life history parameters on organochlorine concentrations in free ranging killer whales. *Sci. Total. Environ*. **281**: 183-203.
- YU, H., BAO, L.J., LIANG, Y. and ZENG, E.Y. 2011. Field validation of anaerobic degradation pathways for dichlorodiphenyltrichloroethane (DDT) and 13 metabolites in marine sediment cores from China. *Environ. Sci. Technol*. **45:** 5245–5252.
- ZAMIR, R., ATHANASIADOU, M., NAHAR, N., MAMUN, M. I. R., MASIHUZZAMAN, M. and BERGMAN, A. 2008. Organohalogen contaminants in plasma from groups of humans with different occupations in Bangladesh. *Chemospher.* **74:** 453-459.
- [ZHAO,](https://www.researchgate.net/profile/Zhonghua_Zhao2) Z., ZHANG, L., [WU, J. and ,](https://www.researchgate.net/profile/Jinglu_Wu) [FAN, C.](https://www.researchgate.net/profile/Chengxin_Fan) 2009. Distribution and bioaccumulation of organochlorine pesticides in surface sediments and benthic organisms from Taihu Lake, China. *Chemospher.* **77(9):**1191-1198.
- ZHOU, H.Y., CHEUNG, R.Y.H. and WONG, M.H. 1999. Residues of organochlorines in sediments and tilapia collected from inland water systems of Hong Kong. *Arch. Environ. Contam. Toxico*. **36:** 424-31.
- ZHOU, R., ZHU, L. and KONG, Q., 2007. Persistent chlorinated pesticides in fish species from Qiantang River in East China. *Chemosphere* **68:** 838–847.
- [ZHOU, R.,](https://www.researchgate.net/profile/Rongbing_Zhou) [ZHU,](https://www.researchgate.net/profile/Rongbing_Zhou) L., [YANG,](https://www.researchgate.net/profile/Kun_Yang42) K. and [CHEN,](https://www.researchgate.net/researcher/38846379_Yuyun_Chen) Y., 2006. Distribution of Organochlorine Pesticides in Surface Water and Sediments from Qiantang River, East China. *J. Hazd. Mater.* **137(1):** 68-75.
- ZHOU, R., ZHU, L., CHEN, Y. and Kong, Q. 2008. Concentrations and characteristics of organochlorine pesticides in aquatic biota from Qiantang River in China. *Environ. Pollut.* **151:** 190–199.
Appendix-1. One way Analysis of Variance (ANOVA) for significance test of DDE, DDD and 2,4-́DDT residues in samples between and within four seasons

ANOVA

ANOVA

DDD

ANOVA

2,4-́ DDT

Appendix-2. One way Analysis of Variance (ANOVA) for significance test of 4,4́-DDT, total DDTs residues and lipid contents in samples between and within four seasons

ANOVA

4,4-́ DDT

ANOVA

ANOVA

Appendix-3. Post Hoc tests (Tukey HSD and LSD) for significance test of DDE residues in samples between and within four seasons

Post Hoc Tests

Multiple Comparisons

Appendix-4. Post Hoc tests (Tukey HSD and LSD) for significance test of DDD residues in samples between and within four seasons

Post Hoc Tests

Multiple Comparisons

Appendix-5. Post Hoc tests (Tukey HSD and LSD) for significance test of 2,4ʹ-DDT residues in samples between and within four seasons

Post Hoc Tests

Multiple Comparisons

Appendix-6. Post Hoc tests (Tukey HSD and LSD) for significance test of 4,4-́DDT residues in samples between and within four seasons

Post Hoc Tests

Multiple Comparisons

Dependent Variable: 4,4ʹ-DDT

Appendix-7. Post Hoc tests (Tukey HSD and LSD) for significance test of total DDTs residues in samples between and within four seasons

Post Hoc Tests

Multiple Comparisons

Appendix-8. Post Hoc tests (Tukey HSD and LSD) for significance test of lipid content in samples between and within four seasons

Post Hoc Tests

Multiple Comparisons

Dependent Variable: Lipid content

Appendix-9. One way Analysis of Variance (ANOVA) for significance test of DDE, DDD and 2,4-́DDT residues in samples between and within different feeding habits

ANOVA

ANOVA

ANOVA

2,4-́ DDT

Appendix-10. One way Analysis of Variance (ANOVA) for significance test of 4,4-́DDT, total DDTs residues and lipid contents in samples between and within different feeding habits

ANOVA

4,4ʹ-DDT

ANOVA

Total DDTs

ANOVA

Appendix-11. Post Hoc tests (Tukey HSD and LSD) for significance test of DDE residues in samples between and within different feeding habits

Post Hoc Tests

Multiple Comparisons

*. The mean difference is significant at the 0.05 level.

Homogeneous Subsets

Means for groups in homogeneous subsets are displayed.

a. Uses Harmonic Mean Sample Size = 22.283.

Appendix-12. Post Hoc tests (Tukey HSD and LSD) for significance test of DDD residues in samples between and within feeding habits

Post Hoc Tests

Multiple Comparisons

*. The mean difference is significant at the 0.05 level.

Homogeneous Subsets

Means for groups in homogeneous subsets are displayed.

a. Uses Harmonic Mean Sample Size = 22.283.

Appendix-13. Post Hoc tests (Tukey HSD and LSD) for significance test of 2,4ʹ-DDT residues in samples between and within different feeding habits

Post Hoc Tests

Multiple Comparisons

Dependent Variable: 2,4ʹ-DDT

Homogeneous Subsets

Means for groups in homogeneous subsets are displayed.

a. Uses Harmonic Mean Sample Size = 22.283.

Appendix-14. Post Hoc tests (Tukey HSD and LSD) for significance test of 4,4ʹ-DDT residues in samples between and within different feeding habits

Post Hoc Tests

Multiple Comparisons

*. The mean difference is significant at the 0.05 level.

Homogeneous Subsets

Means for groups in homogeneous subsets are displayed.

a. Uses Harmonic Mean Sample Size = 22.283.

Appendix-15. Post Hoc tests (Tukey HSD and LSD) for significance test of total DDTs residues in samples between and within different feeding habits

Post Hoc Tests

Multiple Comparisons

*. The mean difference is significant at the 0.05 level.

Homogeneous Subsets

Means for groups in homogeneous subsets are displayed.

a. Uses Harmonic Mean Sample Size = 22.283.

Appendix-16. Post Hoc tests (Tukey HSD and LSD) for significance test of lipid content in samples between and within different feeding habits

Post Hoc Tests

Dependent Variable: Lipid contens

Multiple Comparisons

*. The mean difference is significant at the 0.05 level.

Homogeneous Subsets

Means for groups in homogeneous subsets are displayed.

a. Uses Harmonic Mean Sample Size = 22.283.

Appendix-17. Pearson correlation matrix between DDT and it metabolites, total DDTs, lipid and some meteorological parameters

Correlations

**. Correlation is significant at the 0.01 level (2-tailed).

* Correlation is significant at the 0.05 level (2-tailed)

..

Appendix-18 Hierarchical dendrogram for 20 fishes and prawn represented by DDTs residues obtained by Ward's hierarchical clustering method

Proximities

Case Processing Summary^a

a. Euclidean Distance used

Cluster

Average Linkage (Between Groups)

Hierarchical dendrogram for 20 fishes and prawn represented by DDTs residues obtained by Ward's hierarchical clustering method

Appendix-19 Hierarchical dendrogram for fishes and prawn of summer season represented by DDTs residues obtained by Ward's hierarchical clustering method

Proximities

a. Euclidean Distance used

Cluster

Average Linkage (Between Groups)

Hierarchical dendrogram for fishes and prawn of summer season represented by DDTs residues obtained by Ward's hierarchical clustering method

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